

1 **Worse sleep and increased energy expenditure yet no movement changes in sub-urban**
2 **wild boar experiencing an influx of human visitors (anthropulse) during the COVID-19**
3 **pandemic**

4 Astrid Olejarz^a, Monika Faltusová^a, Luca Börger^b, Justine Guldenpfennig^a, Vilém Jarský^c,
5 Miloš Ježek^a, Euan Mortlock^d, Václav Silovský^a, Tomasz Podgórski^{a,e}

6 ^a Department of Game Management and Wildlife Biology, Faculty of Forestry and Wood
7 Sciences, Czech University of Life Sciences, Kamýcká 129, Prague 6-Suchdol, 165 00, Czech
8 Republic

9 ^b Department of Biosciences, Swansea University, Singleton Park, Swansea SA2 8PP, Wales,
10 UK

11 ^c Department of Forestry and Wood Economics, Faculty of Forestry and Wood
12 Sciences, Czech University of Life Sciences, Kamýcká 129, Prague 6-Suchdol, 165 00, Czech
13 Republic

14 ^d School of Biological Sciences, Queen's University Belfast, 19 Chlorine Gardens, Belfast
15 BT95DL, Northern Ireland

16 ^e Mammal Research Institute, Polish Academy of Sciences, Stoczek 1, 17-230 Białowieża,
17 Poland

18 **ABSTRACT**

19 Expansion of urban areas, landscape transformation and increasing human outdoor activities
20 strongly affect wildlife behaviour. The outbreak of the COVID-19 pandemic in particular led
21 to drastic changes in human behaviour, exposing wildlife around the world to either reduced or
22 increased human presence, potentially altering animal behaviour. Here, we investigate
23 behavioural responses of wild boar (*Sus scrofa*) to changing numbers of human visitors to a
24 suburban forest near Prague, Czech Republic, during the first 2.5 years of the COVID-19
25 epidemic (April 2019 - November 2021). We used bio-logging and movement data of 63 GPS-
26 collared wild boar and human visitation data based on an automatic counter installed in the
27 field. We hypothesised that higher levels of human leisure activity will have a disturbing effect
28 on wild boar behaviour manifested in increased movements and ranging, energy spent, and

29 disrupted sleep patterns. Interestingly, whilst the number of people visiting the forest varied by
30 two orders of magnitude (from 36 to 3431 people weekly), even high levels of human presence
31 (> 2000 visitors per week) did not affect weekly distance travelled, home range size, and
32 maximum displacement of wild boar. Instead, individuals spent 41% more energy at high levels
33 of human presence (> 2000 visitors per week), with more erratic sleep patterns, characterised
34 by shorter and more frequent sleeping bouts. Our results highlight multifaceted effects of
35 increased human activities ('anthropulses'), such as those related to COVID-19
36 countermeasures, on animal behaviour. High human pressure may not affect animal movements
37 or habitat use, especially in highly adaptable species such as wild boar, but may disrupt animal
38 activity rhythms, with potentially detrimental fitness consequences. Such subtle behavioural
39 responses can be overlooked if using only standard tracking technology.

40 **KEYWORDS**

41 Human impact, bio-logging, COVID-19 lockdown, disturbance, *sus scrofa*

42 **1. Introduction**

43 Anthropogenic pressure is growing worldwide, forcing wildlife to adapt to new environmental
44 conditions and human presence (Vitousek et al., 1997, Tuomainen&Candolin 2011; Gunn et al.
45 2022). Expansion of urban areas (Gaynor et al., 2018), habitat fragmentation and landscape
46 transformation (Bruinderink&Hazebroek, 1996; Said et al., 2016; Shi et al., 2018), as well as
47 increasing human outdoor activities (Scholten et al., 2018; Sibbald et al., 2011) affect many
48 aspects of wildlife behaviour. Behavioural responses can include shifts in habitat use and daily
49 activity (Gaynor et al., 2018), overall reduction of movements (Tucker et al. 2018) or diel
50 movements between safe and risky places (Courbin et al. 2022). Wildlife exposed to higher
51 human activity tend to have smaller home ranges and higher rates of social associations at
52 almost all times of the year (Gillich et al., 2021; Grund et al., 2002; Seip et al., 2007).
53 Furthermore, wildlife adjusts its bedding and foraging behaviour in national parks by avoiding

54 hiking or cycling trails during the weekend days with high human visitation rates (Jiang et al.,
55 2007; Scholten et al., 2018; Sibbald et al., 2011), preferring areas that are difficult for humans
56 to reach (Gaynor et al., 2018).

57 The outbreak of the worldwide COVID-19 pandemic at the end of 2019 added yet another
58 dimension to human-wildlife interactions. Epidemic countermeasures, such as restrictions of
59 activity and mobility, led to drastic changes in human behaviour, and with that reduction of
60 disturbance, noise, and other pollution (Bar, 2021). The sudden confinement of roughly two-
61 thirds of the global human population (peak lockdown on April 5, 2020) caused an immediate
62 change in wildlife behaviour (Bates et al., 2020). Shortly after the first implementation of strict
63 lockdowns, social media and online news reported sightings of naturally shy wildlife species in
64 human-occupied landscapes, e.g., pumas in downtown Santiago, Chile or dolphins in the
65 harbour of Trieste, Italy (Max-Planck-Gesellschaft, 2021). Those observations were supported
66 by scientific studies which reported short-term effects of the sudden absence of human pressure,
67 such as an increase of habitat use (Behera et al., 2022), a shift towards diurnal activity (Behera
68 et al., 2022; Manenti et al., 2020; Zukerman et al., 2021), and less roadkill especially of
69 amphibians and reptiles (Driessen, 2021; LeClair et al., 2021; Łopucki et al., 2021; Manenti et
70 al., 2020). On the negative side, an increase in poaching caused by the partial stop of
71 conservation actions was also observed during COVID-19 lockdowns actions (A. E. Bates et
72 al., 2021; Koju et al., 2021; Lindsey et al., 2020; Rahman et al., 2021).

73 Human confinement during the initial COVID-19 lockdowns, termed "anthropause" by Rutz et
74 al. 2020, provided the opportunity to investigate positive and negative effects of human
75 presence and mobility on ecosystems and animal behaviour (Bates et al., 2020). The first
76 COVID-19 lockdowns were followed by a series of periods with relaxed or stringent restrictions
77 depending on the country-specific epidemiological situation. Human mobility fluctuated in
78 accordance with the level of restrictions leading to a series of pulses and pauses of

79 anthropogenic pressure (Rutz, 2022). These COVID-19-related pulses in human activity
80 provide a unique experimental opportunity to test their impacts, yet studies taking such an
81 approach are missing. Government responses to the pandemic varied greatly across the
82 geopolitical spectrum and elicited different responses from the society. Thus, using periods of
83 COVID-19 lockdowns as a simple covariate explaining environmental changes without
84 underlying data on human activity may be insufficient, if not misleading. For example, most
85 reports consider a reduction of human activity during COVID-19 lockdowns, but increased
86 interest in outdoor recreational activities in response to the at-home-confinement was observed
87 in some areas (Hockenhuil et al., 2021; Kleinschroth&Kowarik, 2020; Weed, 2020). Nature
88 parks in particular, where human entry was not restricted, experienced sudden increases in the
89 number of visitors and pressure on the ecosystem. Higher numbers of visitors were observed
90 during lockdown periods (Cukor et al., 2021; Derks et al., 2020; Venter et al., 2020) or shortly
91 after the ease of some restrictions (Day, 2020; McGinlay et al., 2020). For example, in a forest
92 located northeast of the city Zlín in the Czech Republic, the visitation rate of humans in the
93 forest areas increased over five-fold from 200 people per day in April 2019 to 1100 people per
94 day in April 2020 (recorded by 14 randomly placed camera traps), resulting in increased
95 disturbance of wildlife species (Cukor et al., 2021).

96 Whilst many wildlife species are declining due to overexploitation, habitat loss, and traffic
97 mortality. wild boar (*Sus scrofa*) numbers are increasing steadily over the last decades (Massei
98 et al., 2015; Scandura et al., 2021). Studies show that the demographic success of the wild boar
99 is in part due to their high adaptability to a wide range of environmental conditions and
100 tolerance to humans (Fernández-Aguilar et al., 2018). This plasticity enables colonisation of
101 habitats with high human pressure, such as agricultural areas (Morelle et al., 2016), and urban
102 areas (Castillo-Contreras et al., 2018). For example, wild boar shift to nocturnal activity when
103 human presence is high (Boitani et al., 1994; Ikeda et al., 2019; Podgórski et al., 2013; Russo

104 et al., 2010). In response to hunting, wild boar increased movements in search for refuge
105 habitats in dense woodlands to minimise the risk of being detected (Thurfjell et al., 2013).
106 Furthermore, hunting is known to influence the resting behaviour of wild boars. In the period
107 of hunts, the resting areas of the wild boar were clearly larger and more distant from each other
108 (Scillitani et al., 2009; Sodeikat&Pohlmeyer, 2007). Resting areas fulfil an important fitness
109 function for animals, including defence against predators, thermoregulation, rearing of
110 offspring (Lutermann et al., 2010) and sleep. Despite the importance of resting areas, little is
111 known about how increased human presence and activity affects the sleeping behaviour of wild
112 boar.

113 The aim of our study was to describe the effects of changing human presence induced by the
114 countermeasures to COVID-19 pandemic on the movements and space use, activity and sleep,
115 and energy expenditure of wild boar. We hypothesised that higher levels of human leisure
116 activity will have a disturbing effect on wild boar behaviour manifested in increased
117 movements, ranging and energy spent, as well as disrupted sleep patterns. Specifically, we
118 expected to see a positive relationship between weekly number of visitors to the forest and 1)
119 weekly distance travelled, 2) proportion of distance travelled during nighttime (i.e. shift to
120 nocturnality), 3) weekly range size, 4) spatial extent of movements, and 5) energy spent by wild
121 boar. Additionally, we predicted that 1) sleep patterns will become more erratic (shorter and
122 more frequent sleeping bouts) in response to disturbance by high human recreational activity,
123 whereas 2) the total sleep time may remain the same, assuming that recreational activity of
124 people is limited in space (trails) and time (daylight) and thus allow individuals to recover the
125 lost sleep.

126 **2. Material and methods**

127 **2.1. Study area**

128 The study site is located within the municipality "Kostelec nad Černými Lesy", district Prague-
129 East of the Czech Republic (N 49.93` - 49.99`,E 14.72 - 14.88, Figure A.1). The municipality
130 area is covered by 43% of forest, 47% agricultural land, 9% other land-cover types, and 1%
131 water surfaces (Ježek et al., 2016). Our study was conducted in the forested part of the
132 municipality - a 2900 ha woodland administered by the Czech University of Life Sciences
133 Forest Establishment in Kostelec nad Cernymi lesy. The altitude of the study site is 430 m a.s.l.,
134 with a mean annual precipitation of 600mm, and mean annual temperature of 7.5°C (Podrázský
135 et al., 2009). The study area, which offers natural forest landscape and high plant and animal
136 biodiversity, is an attractive place for recreational activities of local and Prague residents
137 (Jarský et al., 2022).

138 **2.2. Wild boar capture and tracking**

139 Wild boars were trapped inside wooden traps using corn as bait. The immobilisation was done
140 by airguns with a mixture of Ketamine, Xylazine and Zoletil inside the darts (Fenati et al.,
141 2008). We followed the protocol of vets and checked the oxygen respiration during the
142 immobilisation of the individuals. The wild boar trapping procedures were in accordance with
143 the decision of the ethics committee of the Ministry of the Environment of the Czech Republic,
144 number MZP/2019/630/361. Captured animals were equipped with hybrid bio-logging collars
145 comprising a GPS unit (Vectronic Aerospace GmbH) and a Daily Diary tag (Wildbyte
146 Technologies Ltd). We recorded biologing data (3–axial accelerometer and 3–axial
147 magnetometer data at 10 Hz frequency) and stored them on the microSD card inside the housing
148 of the Daily Diary. The GPS fixes were collected every 30 minutes and sent by SMS to an
149 online server. We used GPS data of 63 individuals (47 females, 16 males) collected from April
150 2019 to November 2021. For the analysis, we used only GPS fixes with a dilution of precision
151 (DOP) (≥ 1 and ≤ 7) downloaded from the GPS Plus X software, and selected weeks
152 (temporal unit of our study) with at least 5 days of telemetry data with a daily average of at least

153 40 GPS locations. According to these criteria, 135 individual weeks were used for the analyses.
154 Bio-logging data did not cover the study period uniformly and we therefore only used the six
155 most and five least visited weeks for direct comparison. Bio-logging data originated from 13
156 individuals (2 males and 11 females). All GPS data were visualised and analysed using the
157 coordinate reference system EPSG:32633-WGS 84/UTM zone 33N within the R software 4.1.0
158 (R Core Team, 2021).

159 **2.3. Human visitation data**

160 Human presence in the suburban forest was recorded hourly by an automatic counter (eco-
161 counter.com, 2022) at the entrance of the main forest road in Jevany counter (Jarský et al.,
162 2022). We aggregated the human count data into weekly periods, which was the basic temporal
163 unit in our analyses (mean 1126.55 people weekly, 95% confidence interval (CI): 1089.6 –
164 1163.51). There were two COVID-19 lockdown periods during the study period (Figure 1). The
165 lockdowns were defined by the "state of emergency" declared by the government of the Czech
166 Republic (vlada.cz, 2020). The first COVID-19 lockdown in the Czech Republic started on
167 24.03.2020 and ended on 24.04.2020. The second COVID-19 lockdown started on 22.10.2020
168 and ended on 11.04.2021. Furthermore, we divided the study period into seasons: Spring (Mar-
169 May), Summer (Jun-Aug), Autumn (Sep-Nov), and Winter (Dec-Feb) and used season as a
170 covariate.

171 **2.4. Analysis of wild boar movement and space use**

172 Using GPS-telemetry data we calculated the following movement and space use parameters: 1)
173 weekly distance travelled as a sum of all distances between consecutive 30-minute relocations
174 (i.e., step lengths) per week. In addition, we divided the weekly distance into distance travelled
175 at daytime and distance travelled at night time. Daytime was defined from sunrise to sunset and
176 night from sunset to sunrise, 2) weekly home range as 95% kernel utilisation distribution (UD)
177 isopleths using the "reference bandwidth" method from the package "adehabitatHR" (Calenge,

178 2006), 3) maximum displacement as the maximum distance between GPS locations within a
179 week. To examine the effect of human presence on wild boar movement and space use, we used
180 generalised mixed-effects models with the package "lme4" (Bates et al., 2014). In total, we used
181 935 data points (i.e., individual weeks) to fit models to movement and space use data obtained
182 from 63 collared wild boars. For each of the five response variables we fitted a model with fixed
183 effects of weekly human counts (continuous predictor) and season (categorical predictor) as
184 well as animal ID as a random effect. Residuals of all fitted models were normally distributed
185 as evidenced by visual inspection of the quantile plots and histograms of the residuals. The
186 home range and maximum displacement were log-transformed prior to modelling to reduce
187 skewness and improve normality of the residuals. Using the package "ggeffect"(Lüdecke,
188 2018), we generated predictions of the effects of seasons and human activity on wild boar space
189 use and movements in all five models.

190 **2.5. Analysis of wild boar energy expenditure**

191 We used the vectorial sum of dynamic body acceleration (VeDBA) as a proxy for energy
192 expenditure (Wilson et al. 2020). The VeDBA was calculated using the tri-axial acceleration
193 measured by the daily diary tags on the collars. Dynamic body acceleration is a good indicator
194 of oxygen consumption and movement-based power in both humans and animals (Miwa et al.,
195 2017; Qasem et al., 2012; Wilson et al., 2020). We used available biologging data from 12
196 collared wild boars (1 male and 11 female). Using the DDMT software (Wildbyte Technologies
197 Ltd, 2022), we set the smoothing of the VeDBA to 20 records (i.e., 2 seconds) and created 30
198 minute bookmarks. We then exported the sum of the smoothed VeDBA per half an hour for the
199 whole period of available data. However, due to discontinuous data coverage of the study period
200 we selected the top six of the most visited weeks (>2000 visitors) and bottom five weeks of the
201 least visited weeks (<300 visitors; Figure 3), for which data provided by 12 individuals was
202 available. All six weeks that had more than > 2000 visitors per week occurred during the first

203 lockdown. Five weeks with less than < 300 visitors per week occurred during the non-lockdown
204 and the second lockdown. We summarised the smoothed VeDBA for each week using the
205 "collap" package (Krantz et al., 2022) within the R software. This data was obtained from
206 twelve individuals. To examine the differences in VeDBA between the two extreme categories
207 of human visitation, we run a linear mixed model, with the log-transformed VeDBA, human
208 high or low visitation as a fixed effect, and Animal ID as a random effect.

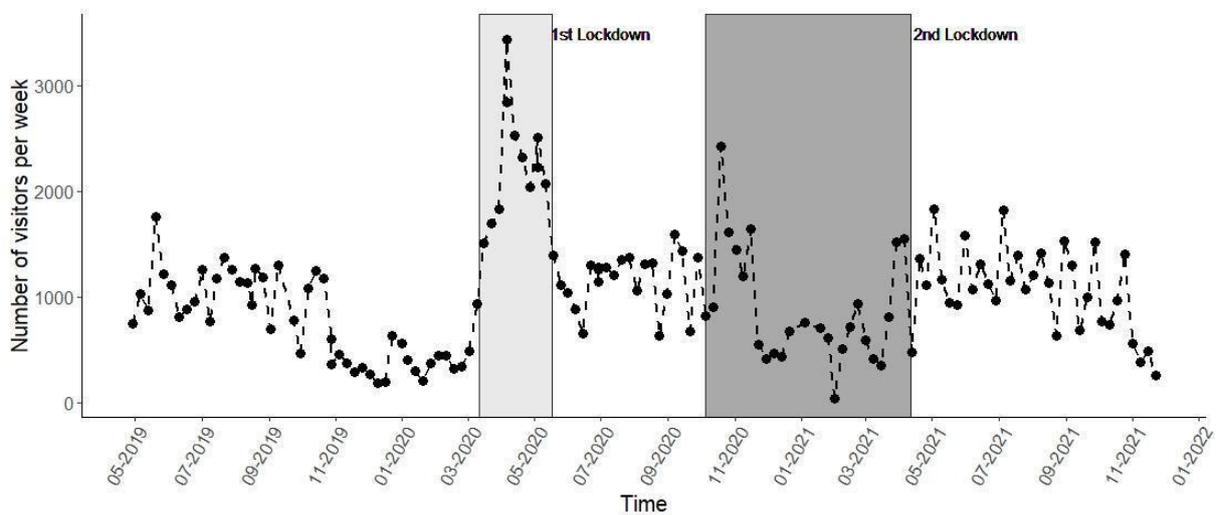
209 **2.6. Analysis of wild boar sleeping behaviour**

210 We used a new method to identify periods of sleep in the daily diary data, developed by
211 modifying existing published laboratory procedures and studies, based on actigraph recordings
212 of sleep in domestic pigs, to use it on accelerometer data collected on wild boar in the wild
213 (Mortlock et al., 2022). Specifically, behavioural sleeping bouts were classified using body
214 pitch and roll angles, identifying the stereotypical sleep postures of either lateral or sternal
215 recumbency, combined with immobility (defined as a VeDBA threshold < 0.2). Furthermore,
216 based on existing physiological measures of sleep in domestic pigs, a transitional period of five
217 minutes was discarded at the start of each bout. After removing the transitional time, the sleep
218 time was calculated. The end of a sleeping bout was identified once the animal started moving,
219 exceeding a smoothed VeDBA threshold of 0.2, which allowed for minor movement during
220 sleep. Using this data, we calculated the average duration of sleep (hours) per animal and day
221 during the specific weeks of high and low human visitation respectively, as well as the number
222 and duration of sleeping bouts as an indicator of sleep continuity within the R software. To
223 examine the differences in the sleeping behaviour between the two extreme categories of human
224 visitation, we run three linear mixed models, with the log-transformed total duration of sleep
225 per week as well as with the number and duration of sleeping bouts as a response variable,
226 human visitation rate (high or low) as a fixed effect, and Animal ID as a random effect.

227 **3. Results**

228 **3.1. Human visitation patterns**

229 We compared human visitation rate obtained from the counter during the two lockdown periods
230 and the non-lockdown period (Kruskal-Wallis chi-squared = 246.09, df = 2, p-value < 0.001).
231 The number of human visitors during the first lockdown (median of 2066 visitors) was
232 significantly higher compared to the second lockdown (902 visitors) and non-lockdown periods
233 (1066 visitors) (pairwise-Wilcox tests, p-value < 0.001). The second lockdown showed no
234 significant difference in the number of visitors compared to the non-lockdown (pairwise-
235 Wilcox test, p = 0.75). Given those results, we believe that the actual visitation rate measured
236 in the field provides better representation of human response to COVID-19 countermeasures
237 than just using the dates of the officially imposed lockdowns. Thus, we used the weekly sum of
238 visitors as a continuous predictor explaining wild boar movements, space use, activity and sleep
239 instead of categorical lockdown and non-lockdown periods.



240
241 **Figure 1: Count of human visitation per week in the forest area near the capital city**
242 **Prague and the two official COVID-19 lockdowns as defined by the "state of emergency"**
243 **declared by the government of the Czech Republic.**

244

245 **3.2. Space-use and movement patterns**

246 We found that the number of visitors in the forest did not affect wild boar spatial behaviour as
 247 none of the five movement parameters was influenced by the weekly human count (Table 1,
 248 Figure A.2). The total weekly distance travelled by wild boar decreased marginally by 145 m
 249 per increase of 400 people visiting the forest and ranged between 34.43 km at 400 visitors and
 250 33.26 km at 3600 visitors (3.4% decrease). The distance travelled during nighttime tended to
 251 decrease while distance travelled during daytime tended to increase when more people visited
 252 the forest (Figure A.2), yet these relationships were statistically insignificant (Table 1). Weekly
 253 home range size was positively, yet insignificantly, related to the number of visitors, showing
 254 a slight increase by 0.26 % per unit of 400 more people visiting the forest. Maximum
 255 displacement was increasing only by 0.06 % per unit of 400 people visiting the forest. Instead,
 256 in contrast to the number of visitors, all five movement and space use parameters varied
 257 significantly across seasons (Table 1).

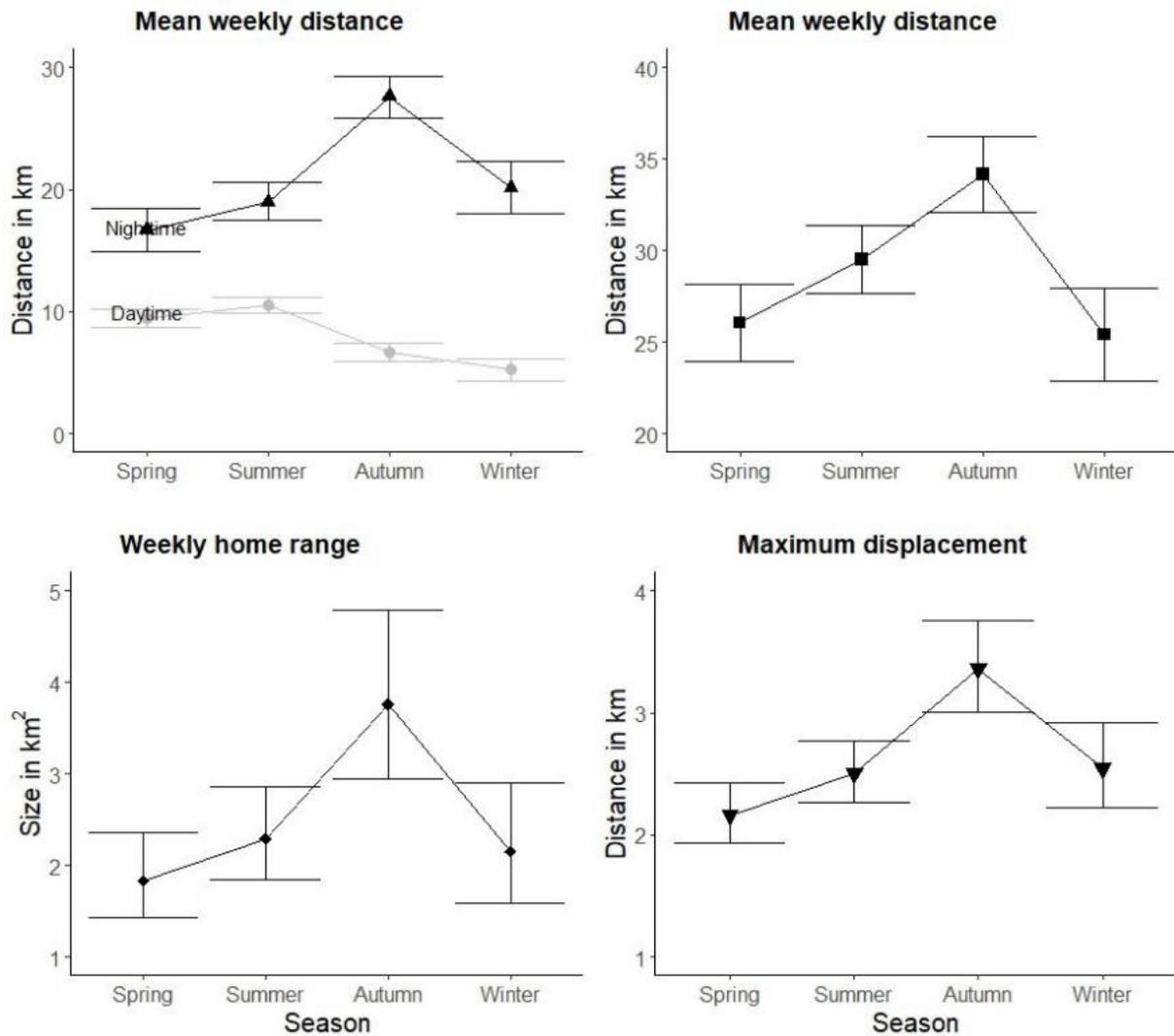
258 **Table 1: Results of the mixed model regression for five estimated movement and space**
 259 **use parameters.**

<i>Coefficient</i>	Weekly daytime distance		Weekly nighttime distance		Weekly home range		Total weekly distance		Maximum displacement	
	<i>Estimates</i>	<i>Conf. Int (95%)</i>	<i>Estimates</i>	<i>Conf. Int (95%)</i>	<i>Estimates</i>	<i>Conf. Int (95%)</i>	<i>Estimates</i>	<i>Conf. Int (95%)</i>	<i>Estimates</i>	<i>Conf. Int (95%)</i>
Autumn (Intercept)	6.18 ***	5.34 – 7.02	28.48 ***	26.54 – 30.42	1.32 ***	1.04 – 1.59	34.57 ***	32.26 – 36.88	1.21 ***	1.09 – 1.34
Human Count	0.00	-0.00 – 0.00	-0.00	-0.00 – 0.00	0.00	-0.00 – 0.00	-0.00	-0.00 – 0.00	-0.00	-0.00 – 0.00
Spring	2.77***	2.04 – 3.49	-10.94 ***	-12.55 – -9.33	-0.72 ***	-0.94 – -0.50	-8.11 ***	-9.95 – -6.26	-0.44 ***	-0.54 – -0.34
Summer	3.89 ***	3.33 – 4.44	-8.61 ***	-9.84 – -7.39	-0.50 ***	-0.66 – -0.33	-4.66 ***	-6.06 – -3.35	-0.29 ***	-0.37 – -0.22
Winter	-1.41 **	-2.26 – -0.57	-7.44 ***	-9.31 – -5.58	-0.56 ***	-0.81 – -0.30	-8.77 ***	-10.92 – -6.62	-0.28 ***	-0.40 – -0.17
Random Effects										
σ^2	9.35		45.54		0.84		59.61		0.17	
τ_{00}	4.69 <i>AnimalID</i>		27.42 <i>AnimalID</i>		0.58 <i>AnimalID</i>		42.17 <i>AnimalID</i>		0.12 <i>AnimalID</i>	
ICC	0.33		0.38		0.41		0.41		0.42	
N	63 <i>AnimalID</i>		63 <i>AnimalID</i>		63 <i>AnimalID</i>		63 <i>AnimalID</i>		63 <i>AnimalID</i>	
Observations	935		935		935		935		934	
Marginal R2 / Conditional R2	0.237 / 0.492		0.201 / 0.501		0.046 / 0.434		0.093 / 0.469		0.078 / 0.461	

260 * $p < 0.05$ ** $p < 0.01$ *** $p < 0.001$

261 Total Weekly distance travelled was highest in autumn (34.17 km on average; CI: 32.11 –
 262 36.22; Figure 2) and lowest in winter (25.40 km on average; CI: 24.48 – 28.02; Figure 2).
 263 Distance travelled at nighttime showed a similar pattern with a peak of 27.61 km (CI: 25.91 –
 264 29.32) in autumn, while the weekly daytime distance peaked in summer at 10.53 km (CI: 9.87

265 – 11.19) and decreased towards winter. Both weekly home range and the maximum
 266 displacement showed similar seasonal patterns with the largest mean values during autumn:
 267 3.76 km² (CI: 2.96 – 4.8) and 3.36km (CI: 3.01 –3.76), respectively (Figure 2).



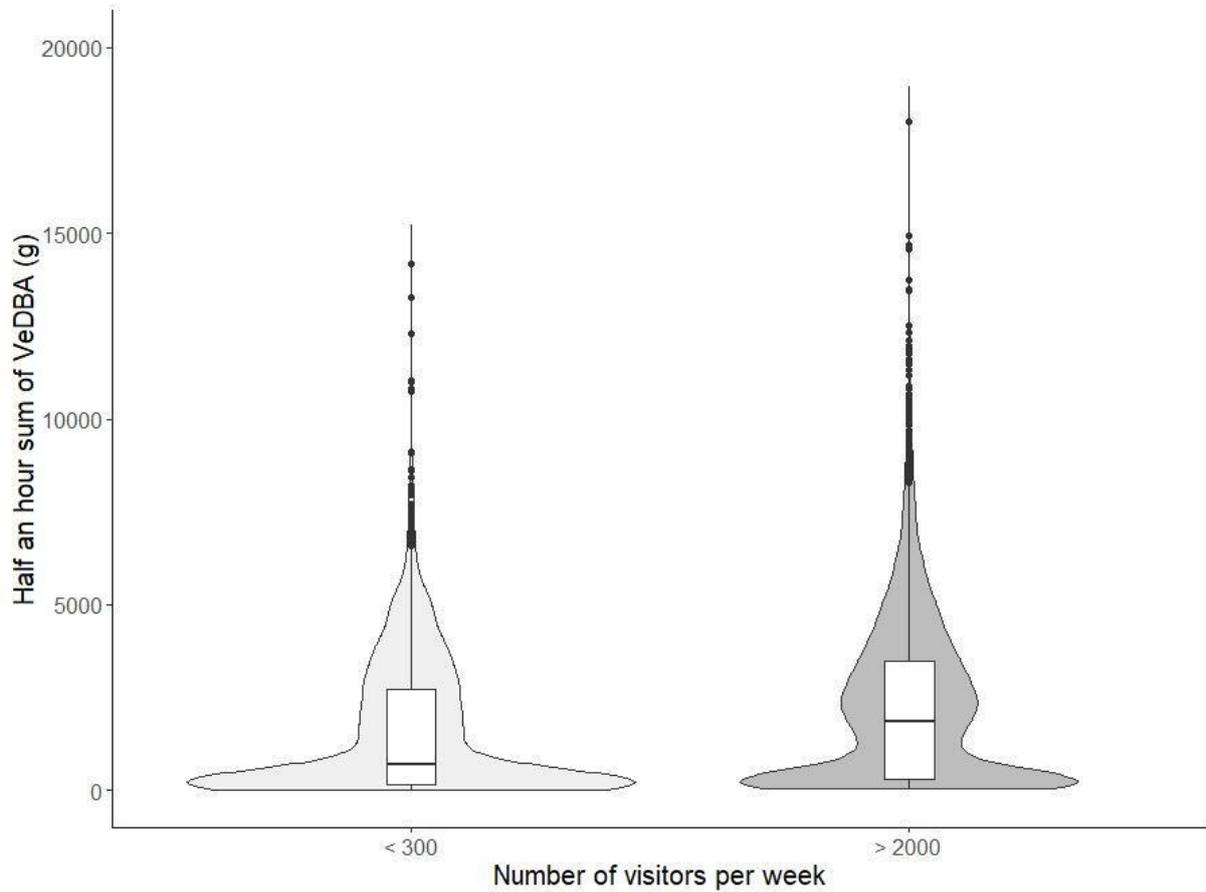
268

269 **Figure2: Seasonal changes in the movement of wild boar: A) Mean weekly distance at**
 270 **nighttime and daytime B) Mean weekly distance C) weekly home range 95 % Kernel D)**
 271 **maximum displacement (maximum distance of GPS locations within a week)**

272 **3.3. Energy expenditure and sleeping behaviour**

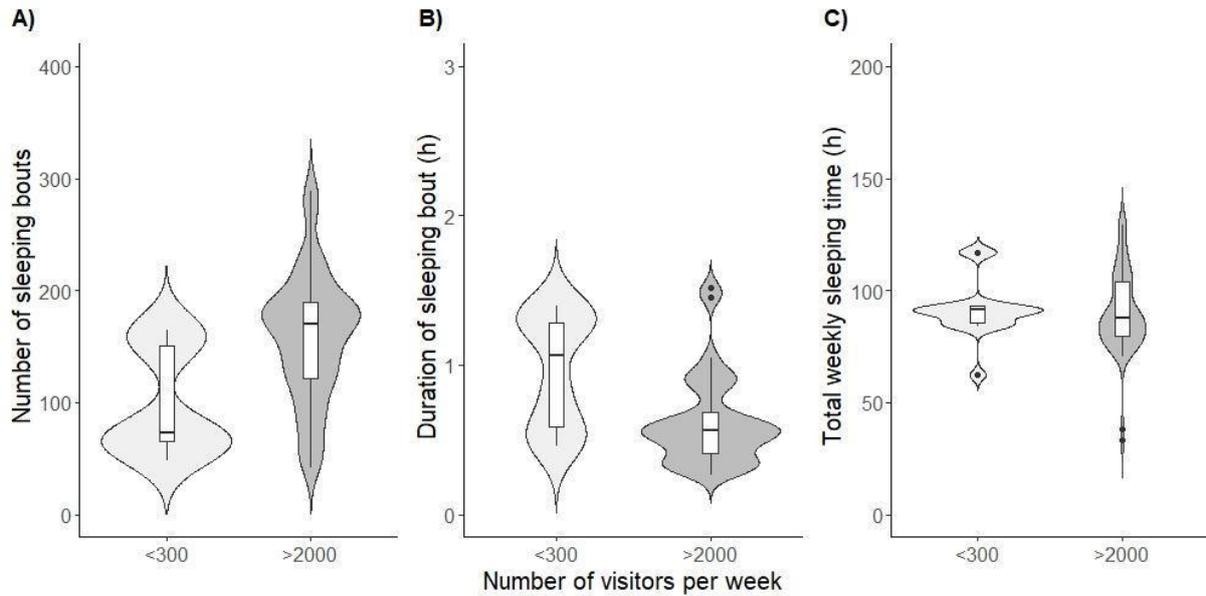
273 The analyses of the wild boar energy expenditure (half an hour sum of VeDBA) showed a 41%
 274 increase in the energy spent between the weeks with the lowest visitation (mean = 1602.24, CI:

275 1529.19 – 1675.3, n = 2448; Figure 3) and the weeks with the highest visitation rates (mean =
276 2260.54, CI: 2216.2 – 2304.7, n = 9215; Figure 3, Table 2).



277
278 **Figure 3: Energy expenditure at the lowest (<300 per week, 5 weeks) and the highest**
279 **(>2000 per week, 6 weeks) number of human visitors.**

280 Total weekly sleep time did not differ much between weeks with high (mean = 90.53 hours per
281 week, CI: 88.08 – 92.97, n = 212) and low human visitor numbers (mean = 91.41, CI: 87.9 –
282 94.93, n = 51; Figure 4, Table 2). However, we observed significantly more sleeping bouts
283 during weeks with high human visitation (mean = 161.63, CI: 154.19 – 169.07, n = 212) than in
284 weeks with few visits (mean = 102.4, CI: 89.52 – 115.26, n = 51; Figure 4; Table 2).
285 Accordingly, the average duration of a sleeping bout was shorter with high human visitation
286 (mean = 0.64 h, CI: 0.602 – 0.684, n = 212) than in weeks with few visits in the forest (mean =
287 0.98 h, CI = 0.874 – 1.09, n = 51; Figure 4, Table 2).



288

289 **Figure 4: Sleeping behaviour at the lowest (<300 per week, 5 weeks) and the highest**
 290 **(>2000 per week, 6 weeks) numbers of human visitors: A) number of sleeping bouts per**
 291 **week B) duration of sleeping bouts C) total sleeping time per week.**

292 Except for the analysis of the total sleep time, linear mixed models of the weekly energy
 293 expenditure, number of sleeping bouts and duration of sleep bouts showed a significant
 294 difference between weeks with low and high human visitation (Table 2).

295 **Table 2: Results of the mixed model regression for sleep metrics and energy expenditure.**

Coefficient	Number weekly sleep bouts		Duration weekly sleep bouts		Total weekly sleep time		Weekly energy expenditure	
	Estimates	Conf. Int (95%)	Estimates	Conf. Int (95%)	Estimates	Conf. Int (95%)	Estimates	Conf. Int (95%)
Visitation < 300 (Intercept)	4.81***	4.58 – 5.05	-0.44***	-0.63 – -0.25	4.49***	4.39 – 4.59	1579.79***	1161.33 – 1998.25
Visitation > 2000	0.11*	0.00 – 0.22	-0.13**	-0.23 – -0.04	-0.01	-0.09 – 0.07	626.31*	83.59 – 1169.03
Random Effects								
σ^2	0.05		0.04		0.03		4338347.31	
τ_{00}	0.17	AnimalID	0.11	AnimalID	0.02	AnimalID	213393.27	AnimalID
ICC	0.77		0.76		0.40		213393.27	
N	13	AnimalID	13	AnimalID	13	AnimalID	12	AnimalID
Observations	287		287		287		11663	
Marginal R ² / Conditional R ²	0.009 / 0.772		0.019 / 0.762		0.000 / 0.401		0.014 / 0.060	

* p<0.05 ** p<0.01 *** p<0.001

296

297 4. Discussion

298 4.1. Human presence during COVID-19 lockdown

299 We showed that the numbers of human visitors to the suburban forest "Kostelec nad cernymi
 300 lesy" of Prague and hence the intensity of recreational use of the forest varied markedly between

301 the two Covid-19 lockdowns. During the first COVID-19 lockdown, there was a strong increase
302 in visitors to the study area which exceeded all levels recorded during the pre-lockdown period
303 as well as those recorded in the following year. This effect can be explained by the type of
304 restrictions imposed on school, work, and recreational facilities by the government during the
305 "state of emergency" declared in the Czech Republic to deal with the Covid-19 pandemic.
306 During this first lockdown, natural areas, parks, and forests were one of the few places freely
307 accessible for visitors and they attracted people seeking relief from the at-home-confinement.
308 Contrastingly, the number of visitors to the forest did not increase during the second lockdown.
309 Although the "state of emergency" was declared in both lockdowns, the restrictions in the
310 second lockdown were much more severe in addition to the restrictions on school, work and
311 recreational facilities, further restrictions on travelling between municipalities (prohibited
312 under a penalty of a fine) were implemented and a curfew was imposed between 9pm and 6am.
313 Those additional restrictions likely discouraged people from extended travelling and made
314 forest visits less likely. Patterns of fluctuating human pressure (i.e., anthropulses) observed in
315 our study highlight the need of using the actual indices of human activity rather than crude
316 administrative measures (i.e., timing of lockdowns or state of emergency declaration) because
317 small changes in the details of each policy can have profound effects on human behaviour and
318 potentially on wildlife.

319 **4.2.Human disturbance and wild boar movement**

320 During our study, human visitation rate in study area fluctuated greatly (varying by two orders
321 of magnitude), yet we did not detect any significant difference in space use and movement
322 patterns of wild boar resulting from these changes. This agrees with the high tolerance and
323 habituation towards anthropogenic pressure recorded for wild boars in urban areas (Licoppe et
324 al., 2013). Similarly, urban wild boars are characterised by a shorter flight distance and reuse
325 of traps (Stillfried et al., 2017). We suspect that the suburban forest is exposed to a constant

326 high pressure of human leisure activities, so that behavioural response of wild boar to human
327 presence may already have occurred before the sharp increase in visitor numbers during the
328 first lockdown. This is supported by our observation of larger distances travelled by wild boar
329 at nighttime across seasons, in accordance with several studies reporting more nocturnal activity
330 of wild boar in response to human disturbances (Gaynor et al., 2018; Johann, et al., 2020a;
331 Podgórski et al., 2013). Hunting events, depending on location and type, can cause instability
332 in wild boar spatio-temporal behaviour but the effects vary across studies (Keuling et al 2021).
333 Some publications report an increase of home range size (Scillitani et al 2009), whilst others
334 report a spatial shift of home range after hunts (Sodeikat&Pohlmeyer 2002, 2003) or did not
335 observe any significant change in home range size (Keuling et al 2008b). Conversely, our
336 results indicate that non-lethal human leisure activities, which are usually restricted to
337 established roads and paths, may not be as disturbing as hunts, and thus do not lead to temporal
338 displacement of animals. Our findings provide similar conclusions to Fatterbert et al (2017)
339 who found that non-lethal human disturbances, measured by the proximity to infrastructures, in
340 the Geneva Basin, Switzerland, had no effect on wild boar ranging patterns. In addition, whilst
341 landscape configuration and topography can have a strong effect on the home range size of wild
342 boar (Fatterbert et. al 2017), our study area was relatively homogenous in terms of forest
343 configuration (continuous cover) and topography (minor differences in elevation), and we did
344 not consider those variables a strong drivers of wild boar spatial behaviour.

345 **4.3. Seasonal effects on wild boar movement**

346 Contrary to the effect of human presence, we found a strong seasonal effect on all our movement
347 and space use parameters, suggesting that wild boar movements and space use are more strongly
348 affected by the species annual life cycle or by resource distribution than by human leisure
349 activities. Weekly distance travelled, weekly home range and maximum displacement showed
350 a similar seasonal pattern with the highest values observed in autumn. As a capital breeder,

351 gaining sufficient fat reserves before winter is crucial for wild boar survival and reproduction
352 in the following year (Geisser&Reyer, 2005; Jędrzejewska et al., 1997). The autumn mast of
353 oak acorn and beech nuts provides natural resources to achieve good body condition before
354 winter but localising those resources may require extended movements and higher spatial
355 activity. Additionally, during the mating season (October - December, Rosell et al., 2012), male
356 wild boar roam widely and often undertakes mating excursions outside of their home range in
357 search of receptive females (Singer et al., 1981), which could further explain the increased home
358 range sizes observed in autumn. In winter, home ranges can increase due to food shortage
359 (Boitani et al., 1994) but not after a tree masting season (Keuling et al., 2008a). We did not
360 observe any home range size increase during the winter period, possibly due to the
361 supplementary feeding practised by managers in the study area. The smallest weekly home
362 ranges were observed during spring which coincides with the peak of parturition and weaning
363 of newborn piglets, whereas in early summer the increasing movement capacity of growing
364 piglets, and high energy demands of sows still nursing the piglets result in larger home ranges
365 compared to spring (Keuling et al., 2008b). As our dataset was female-biased and these seasonal
366 changes in female behaviour may have particularly affected the seasonal space use patterns we
367 observed. Finally, weather conditions can also strongly influence animal movement
368 behaviour in addition to regular seasonal changes (Börger et al., 2006). The more extreme the
369 weather is, the less wild boar move; in winter snow depth and low temperature can reduce the
370 movement activity of wild boar (Johann et al., 2020b; Thurfjell et al., 2014), as do high
371 temperatures in summer (Johann, et al., 2020a).

372 **4.4. Effect of human disturbance on wild boar energy expenditure and sleeping behaviour**

373 Increased human presence on roads and trails in the suburban forest significantly affected the
374 index of energy expenditure (VeDBA) of wild boar. It was 41% higher in the weeks where more
375 than 2000 visitors were counted in the forest than in the weeks with less than 300 visitors. Taken

376 together, our results show that higher recreational human activity did not cause an increase in
377 travel distances, as could be expected for a species habituated to human presence, but
378 sufficiently disturbed the individuals to cause an increase in small-scale body movements and
379 activity on site, as evidenced by higher energy expenditure values. Typically, at high human
380 disturbance levels, wild boars spend their daytime resting in forests and dense shrubbery areas
381 (Boitani et al., 1994). However, at extreme values of human presence (>2000 visitors), animals
382 may have trouble finding sufficiently secluded resting sites and may need to increase their
383 vigilance and thus energy expenditure. Small on-site movements (i.e. non-travel), not detectable
384 by the 30-minute scale GPS data, may also have occurred, but importantly these did not lead to
385 the individuals moving away from their sites (which would have been detected by the GPS
386 data).

387 Our analyses of sleep patterns at high and low human visitation rate further support this
388 prediction. Wild boar sleep was more fragmented (short and frequent sleeping bouts) when
389 human presence on forest roads was high compared to weeks of low human presence, where
390 sleep was more consolidated and thus of higher quality (longer but fewer bouts of sleep).
391 Despite the differences in sleep pattern, total sleep time was similar at high and low human
392 visitation rate. The total sleep time of wild boars may not be affected by human presence.
393 Instead, environmental conditions, such as temperature, humidity, precipitation and snow cover
394 can affect both sleep duration and structure in wild boar (Mortlock et al., 2022). Sleep quantity
395 and quality also varies across and within individuals (Mortlock et al., 2022), which may help
396 explain high variability in the weekly sleep measures observed in our study. Sleep,
397 characterised by rest and reduced reactivity (Zaid et al., 2022), has fundamental functions for
398 the immune (Rogers et al., 2001), neuronal (McDermott et al., 2003) and cognitive system
399 (Roth et al., 2010) in all animals in which sleep has been recorded. Depending on the species,
400 sleep quality differs in duration and number of sleeping bouts during the day (Capellini et al.,

401 2008). Elephants, for example, need only a small amount of sleep, an average daily total sleep
402 time of 2h being enough (Gravett et al., 2017). In contrast, the total daily sleep duration of a
403 sloth is between 9 and 10 hours (Voirin et al., 2014). Sleep is so essential that lack of sleep can
404 be fatal for the animal (Rechtschaffen and Bergmann, 2002). Although sleep fragmentation
405 does not necessarily reduce the total sleep time, as in our study, it has an impact on the sleep
406 quality (Martin et al., 2012) and may negatively impact metabolic stability or endocrine and
407 autonomous systems (Baud et al., 2013). Fragmentation of sleep can cause increased sleepiness,
408 decreased psychomotor performance such as reduced short-term memory, reaction time, or
409 vigilance (Bonnot and Arand et al., 2003, Phillipson et al., 1980). Further, in humans sleep
410 disturbance negatively affects cardiovascular health (Gangwisch et al., 2005). Social and
411 ecological pressures, such as predation risk, food competition, and social relationships, can
412 influence sleep homeostasis in animals (Loftus et al., 2022, Voirin et al., 2014). Within the
413 context of sleep, our results provide new evidence that short-term increased leisure human
414 activity can disrupt sleep quality in a natural setting even in a species with high tolerance to
415 human presence like the wild boar. Our high-resolution approach to quantifying sleep allowed
416 us to see that although wild boar sleep duration was unaffected, sleep quality was reduced by
417 disturbance (being more fragmented), highlighting the need for ecologists to view sleep
418 behaviour in multiple dimensions to capture all potential effects. Our findings are therefore
419 important for the management of natural areas, in particular of eco-tourism and use of green
420 areas by humans. If high numbers of humans visiting natural areas are maintained over
421 prolonged periods, this may have a cumulative deleterious effect on animal physiology and
422 survival. The consequences of sleep disturbance and deprivation in wild animals is a topic
423 requiring further study, holding significance for management and conservation of wildlife
424 populations in human-dominated landscapes.

425 **4.5. Conclusions**

426 Our results show that high levels of human recreational activity, mostly restricted to tourist
427 trails and forest roads, did not affect wild boar space use and long-distance movements.
428 However, we showed that increased human presence influenced *in situ* body movements and
429 sleep behaviour. Disrupted sleeping behaviour, identified as increased sleep fragmentation,
430 could lead to increased energy expenditure and elevated stress levels and disrupt the vital
431 functions of sleep in maintaining natural immunity and neuronal and cognitive functions
432 (Ferrara & De Gennaro, 2001; Rogers et al., 2001) with potentially serious consequences on
433 fitness. We thus highlight the need for more detailed research on the effects of non-lethal human
434 disturbance on animal behaviour to better manage human-wildlife coexistence.

435 **Data availability**

436 The dataset analysed during the current study is available on request to the corresponding
437 authors.

438 **CRediT authorship contribution statement**

439 **A. Olejarz:** Conceptualisation, Methodology, Data processing and analysis, Writing. **M.**
440 **Faltusová:** Conceptualization. **L. Börger:** Data processing and analysis, Writing **J.**
441 **Güldenpfennig:** Methodology, Data analysis, Writing. **V Jarský:** Data collection. **M. Ježek:**
442 Data collection. **E. Mortlock:** Data processing and analysis. **V. Silovský:** Data collection and
443 processing, Writing. **T. Podgorski:** Conceptualisation, Supervision, Data analysis, Writing.

444 **Declaration of competing interest**

445 The authors declare that they have no conflict of interest.

446 **Ethical approval**

447 The wild boar trapping was realised in accordance with the decision of the ethics committee of
448 the Ministry of the Environment of the Czech Republic number MZP/2019/630/361.

449 **Acknowledgements**

450 This work was supported by the University Grant Competition at the Czech University of Life
451 Sciences in Prague No. 82/2021; OP RDE project Improvement in Quality of the Internal Grant
452 Scheme at CZ, (No.CZ,.02.2.69/0.0/0.0/19_073/0016944); "EVA4.0" grant
453 (No.CZ,.02.1.01/0.0/0.0/16_019/0000803); OP RDE and grant (No. QK1910462) financed by
454 the Ministry of Agriculture of the Czech Republic.

455 **References**

- 456 Bar, H. (2021). COVID-19 lockdown: Animal life, ecosystem and atmospheric environment.
457 *Environment, Development and Sustainability*, 23(6), 8161–8178.
458 <https://doi.org/10.1007/s10668-020-01002-7>
- 459 Bates, A. E., Primack, R. B., Biggar, B. S., Bird, T. J., Acevedo-Charry, O., Colón-Piñero, Z., Ocampo, D.,
460 Ocampo-Peñuela, N., Sánchez-Clavijo, L. M., Adamescu, C. M., Cheval, S., Racoviceanu, T.,
461 Kuire, V. Z., Aditya, V., Anderwald, P., Wiesmann, S., Wipf, S., Badihi, G., Henderson, M. G., ...
462 Duarte, C. M. (2021). Global COVID-19 lockdown highlights humans as both threats and
463 custodians of the environment. *Biological Conservation*, 263, 109175.
464 <https://doi.org/10.1016/j.biocon.2021.109175>
- 465 Bates, A. E., Primack, R. B., Moraga, P., & Duarte, C. M. (2020). COVID-19 pandemic and associated
466 lockdown as a “Global Human Confinement Experiment” to investigate biodiversity
467 conservation. *Biological Conservation*, 248, 108665.
468 <https://doi.org/10.1016/j.biocon.2020.108665>
- 469 Bates, D., Mächler, M., Bolker, B., & Walker, S. (2014). Fitting Linear Mixed-Effects Models Using lme4.
470 ArXiv E-Prints, arXiv:1406. <https://doi.org/10.18637/jss.v067.i01>
- 471 Baud, M.O., Magistretti, P.J., Petit, J.-M., 2013. Sustained sleep fragmentation affects brain
472 temperature, food intake and glucose tolerance in mice. *Journal of Sleep Research* 22, 3–12.
- 473 Behera, A. K., Kumar, P. R., Priya, M. M., Ramesh, T., & Kalle, R. (2022). The impacts of COVID-19
474 lockdown on wildlife in Deccan Plateau, India. *Science of The Total Environment*, 822, 153268.
475 <https://doi.org/10.1016/j.scitotenv.2022.153268>
- 476 Boitani, L., Mattei, L., Nonis, D., & Corsi, F. (1994). Spatial and Activity Patterns of Wild Boars in Tuscany,
477 Italy. *Journal of Mammalogy*, 75(3), 600–612. <https://doi.org/10.2307/1382507>
- 478 Bonnet, M.H., Arand, D.L., 2003. Clinical effects of sleep fragmentation versus sleep deprivation. *Sleep*
479 *Medicine Reviews* 7, 297–310.
- 480 Börger, L., Franconi, N., Ferretti, F., Meschi, F., Michele, G., Gantz, A., Coulson, T., 2006. An Integrated
481 Approach to Identify Spatiotemporal and Individual-Level Determinants of Animal Home
482 Range Size. *The American naturalist* 168, 471–85.
- 483 Bruinderink, G. W. T. A. G., & Hazebroek, E. (1996). Ungulate Traffic Collisions in Europe. *Conservation*
484 *Biology*, 10(4), 1059–1067. <https://doi.org/10.1046/j.1523-1739.1996.10041059.x>
- 485 Calenge, C. (2006). The package “adehabitat” for the R software: A tool for the analysis of space and
486 habitat use by animals. *Ecological Modelling*, 197(3), 516–519.
487 <https://doi.org/10.1016/j.ecolmodel.2006.03.017>
- 488 Capellini, I., Nunn, C.L., McNamara, P., Preston, B.T., Barton, R.A., 2008. Energetic constraints, not
489 predation, influence the evolution of sleep patterning in mammals. *Functional Ecology* 22,
490 847–853.
- 491 Castillo-Contreras, R., Carvalho, J., Serrano, E., Mentaberre, G., Fernández-Aguilar, X., Colom, A.,
492 González-Crespo, C., Lavín, S., & López-Olvera, J. R. (2018). Urban wild boars prefer fragmented
493 areas with food resources near natural corridors. *Science of The Total Environment*, 615, 282–
494 288. <https://doi.org/10.1016/j.scitotenv.2017.09.277>

495 Courbin, N., Garel, M., Marchand, P., Duparc, A., Debeffe, L., Börger, L., Loison, A., 2022. Interacting
496 lethal and nonlethal human activities shape complex risk tolerance behaviors in a mountain
497 herbivore. *Ecological Applications* 32, e2640.

498 Cukor, J., Linda, R., Mahlerová, K., Vacek, Z., Faltusová, M., Marada, P., Havránek, F., & Hart, V. (2021).
499 Different patterns of human activities in nature during Covid-19 pandemic and African swine
500 fever outbreak confirm direct impact on wildlife disruption. *Scientific Reports*, 11(1), 1.
501 <https://doi.org/10.1038/s41598-021-99862-0>

502 Day, B. H. (2020). The Value of Greenspace Under Pandemic Lockdown. *Environmental and Resource*
503 *Economics*, 76(4), 1161–1185. <https://doi.org/10.1007/s10640-020-00489-y>

504 Derks, J., Giessen, L., & Winkel, G. (2020). COVID-19-induced visitor boom reveals the importance of
505 forests as critical infrastructure. *Forest Policy and Economics*, 118, 102253.
506 <https://doi.org/10.1016/j.forpol.2020.102253>

507 Driessen, M. M. (2021). COVID-19 restrictions provide a brief respite from the wildlife roadkill toll.
508 *Biological Conservation*, 256, 109012. <https://doi.org/10.1016/j.biocon.2021.109012>

509 Eco-Counter. (2022). Mobile MULTI - Mobile/temporary bike & pedestrian counter. Eco-Counter.
510 <https://www.eco-counter.com/produits/multi-range/mobile-multi>

511 Fattebert, J., Baubet, E., Slotow, R., Fischer, C., 2017. Landscape effects on wild boar home range size
512 under contrasting harvest regimes in a human-dominated agro-ecosystem. *Eur J Wildl Res* 63,
513 32 <https://doi.org/10.1007/s10344-017-1090-9>

514 Fenati, M., Monaco, A., & Guberti, V. (2008). Efficiency and safety of xylazine and tiletamine/zolazepam
515 to immobilize captured wild boars (*Sus scrofa* L. 1758): Analysis of field results. *European*
516 *Journal of Wildlife Research*, 54(2), 269–274. <https://doi.org/10.1007/s10344-007-0140-0>

517 Fernández-Aguilar, X., Gottschalk, M., Aragon, V., Càmara, J., Ardanuy, C., Velarde, R., Galofré-Milà, N.,
518 Castillo-Contreras, R., López-Olvera, J. R., Mentaberre, G., Colom-Cadena, A., Lavín, S.,
519 & Cabezón, O. (2018). Urban Wild Boars and Risk for Zoonotic *Streptococcus suis*, Spain.
520 *Emerging Infectious Diseases*, 24(6), 1083–1086. <https://doi.org/10.3201/eid2406.171271>

521 Ferrara, M., & De Gennaro, L. (2001). How much sleep do we need? *Sleep Medicine Reviews*, 5(2), 155–
522 179. <https://doi.org/10.1053/smr.2000.0138>

523 Gangwisch, J. E., Malaspina, D., Boden-Albala, B., & Heymsfield, S. B. (2005). Inadequate sleep as a risk
524 factor for obesity: Analyses of the NHANES I. *Sleep*, 28(10), 1289–1296.
525 <https://doi.org/10.1093/sleep/28.10.1289>

526 Gaynor, K. M., Hojnowski, C. E., Carter, N. H., & Brashares, J. S. (2018). The influence of human
527 disturbance on wildlife nocturnality. *Science*, 360(6394), 1232–1235.
528 <https://doi.org/10.1126/science.aar7121>

529 Geisser, H., & Reyer, H.-U. (2005). The influence of food and temperature on population density of wild
530 boar *Sus scrofa* in the Thurgau (Switzerland). *Journal of Zoology*, 267, 89–96.
531 <https://doi.org/10.1017/S095283690500734X>

532 Gillich, B., Michler, F.-U., Stolter, C., & Rieger, S. (2021). Differences in social-space–time behaviour of
533 two red deer herds (*Cervus elaphus*). *Acta Ethologica*. <https://doi.org/10.1007/s10211-021-00375-w>

535 Gravett, N., Bhagwandin, A., Sutcliffe, R., Landen, K., Chase, M.J., Lyamin, O.I., Siegel, J.M., Manger,
536 P.R., 2017. Inactivity/sleep in two wild free-roaming African elephant matriarchs – Does large
537 body size make elephants the shortest mammalian sleepers? *PLOS ONE* 12, e0171903.

538 Grund, M. D., McAninch, J. B., & Wiggers, E. P. (2002). Seasonal Movements and Habitat Use of Female
539 White-Tailed Deer Associated with an Urban Park. *The Journal of Wildlife Management*, 66(1),
540 123–130. <https://doi.org/10.2307/3802878>

541 Gunn, R.L., Hartley, I.R., Algar, A.C., Niemelä, P.T., Keith, S.A., 2022. Understanding behavioural
542 responses to human-induced rapid environmental change: a meta-analysis. *Oikos* 2022,
543 e08366.

544 Gunner, R. M., Holton, M. D., Scantlebury, D. M., Hopkins, P., Shepard, E. L. C., Fell, A. J., Garde, B.,
545 Quintana, F., Gómez-Laich, A., Yoda, K., Yamamoto, T., English, H., Ferreira, S., Govender, D.,
546 Viljoen, P., Bruns, A., van Schalkwyk, O. L., Cole, N. C., Tatayah, V., ... Wilson, R. P. (2021). How

547 often should dead-reckoned animal movement paths be corrected for drift? *Animal*
548 *Biotelemetry*, 9(1), 43. <https://doi.org/10.1186/s40317-021-00265-9>

549 Hockenhull, J., Squibb, K., & Cameron, A. (2021). How Has the COVID-19 Pandemic Affected the Way
550 We Access and Interact with the Countryside and the Animals within It? *Animals: An Open*
551 *Access Journal from MDPI*, 11(8), 2281. <https://doi.org/10.3390/ani11082281>

552 Ikeda, T., Kuninaga, N., Suzuki, T., Ikushima, S., & Suzuki, M. (2019). Tourist-wild boar (*Sus scrofa*)
553 interactions in urban wildlife management. *Global Ecology and Conservation*, 18, e00617.
554 <https://doi.org/10.1016/j.gecco.2019.e00617>

555 Jarský, V., Palátová, P., Riedl, M., Zahradník, D., Rinn, R., & Hochmalová, M. (2022). Forest Attendance
556 in the Times of COVID-19-A Case Study on the Example of the Czech Republic. *International*
557 *Journal of Environmental Research and Public Health*, 19(5), 2529.
558 <https://doi.org/10.3390/ijerph19052529>

559 Jędrzejewska, B., Jędrzejewski, W., Bunevich, A. N., Miłkowski, L., & Krasieński, Z. A. (1997). *Factors*
560 *shaping population densities and increase rates of ungulates in Białowieża Primeval Forest*
561 *(Poland and Belarus) in the 19th and 20th centuries*. 53.

562 Ježek, M., Holá, M., Tomáš, K., & Jaroslav, Č. (2016). Creeping into a wild boar stomach to find traces
563 of supplementary feeding. *Wildlife Research*, 43, 590–598. <https://doi.org/10.1071/WR16065>

564 Jiang, G., Minghai, Z., & Ma, J. (2007). Effects of human disturbance on movement, foraging and bed
565 selection in red deer *Cervus elaphus xanthopygus* from the Wandashan Mountains,
566 northeastern China. *Acta Theriologica*, 52, 435–446. <https://doi.org/10.1007/BF03194241>

567 Johann, F., Handschuh, M., Linderoth, P., Dormann, C. F., & Arnold, J. (2020). Adaptation of wild boar
568 (*Sus scrofa*) activity in a human-dominated landscape. *BMC Ecology*, 20(1), 4.
569 <https://doi.org/10.1186/s12898-019-0271-7>

570 Johann, F., Handschuh, M., Linderoth, P., Heurich, M., Dormann, C., & Arnold, J. (2020). Variability of
571 daily space use in wild boar *Sus scrofa*. *Wildlife Biology*, 2020.
572 <https://doi.org/10.2981/wlb.00609>

573 Keuling, O., Stier, N., Roth, M., 2008a. Annual and seasonal space use of different age classes of female
574 wild boar *Sus scrofa* L. *European Journal of Wildlife Research* 54, 403–
575 412 <https://doi.org/10.1007/s10344-007-0157-4>.

576 Keuling, O., Stier, N., Roth, M., 2008b. How does hunting influence activity and spatial usage in wild
577 boar *Sus scrofa* L.? *Eur J Wildl Res* 54, 729 <https://doi.org/10.1007/s10344-008-0204-9>

578 Keuling, O., Massei, G., 2021. Does Hunting Affect the Behavior of Wild Pigs? *Human–Wildlife*
579 *Interactions* 15 <https://doi.org/10.26077/3a83-9155>

580 Kleinschroth, F., & Kowarik, I. (2020). COVID-19 crisis demonstrates the urgent need for urban
581 greenspaces. *Frontiers in Ecology and the Environment*, 18(6), 318–319.
582 <https://doi.org/10.1002/fee.2230>

583 Koju, N. P., Kandel, R. C., Acharya, H. B., Dhakal, B. K., & Bhujju, D. R. (2021). COVID-19 lockdown frees
584 wildlife to roam but increases poaching threats in Nepal. *Ecology and Evolution*, 11(14), 9198–
585 9205. <https://doi.org/10.1002/ece3.7778>

586 Krantz, S., Dowle, M., Srinivasan, A., Jacob, M., Eddelbuettel, D., Berge, L., Tappe, K., worldwide, R. C.
587 T. and contributors, Plummer, M., & Team, 1999-2016 The R. Core. (2022). *collapse: Advanced*
588 *and Fast Data Transformation* (1.8.6) [Computer software]. [https://CRAN.R-](https://CRAN.R-project.org/package=collapse)
589 [project.org/package=collapse](https://CRAN.R-project.org/package=collapse)

590 LeClair, G., Chatfield, M. W. H., Wood, Z., Parmelee, J., & Frederick, C. A. (2021). Influence of the COVID-
591 19 pandemic on amphibian road mortality. *Conservation Science and Practice*, 3(11), e535.
592 <https://doi.org/10.1111/csp2.535>

593 Licoppe, A., Prévot, C., Heymans, M., Bovy, C., Casaer, J., Cahill, S., 2013. Wild boar/feral pig in (peri-)
594 urban areas. In: *Managing Wild Boar in Human-Dominated Landscapes*. International Union of
595 Game Biologists—Congress IUGB. pp. 1–31.

596 Lindsey, P., Allan, J., Brehony, P., Dickman, A., Robson, A., Begg, C., Bhammar, H., Blanken, L., Breuer,
597 T., Fitzgerald, K., Flyman, M., Gandiwa, P., Giva, N., Kaelo, D., Nampindo, S., Nyambe, N.,
598 Steiner, K., Parker, A., Roe, D., ... Tyrrell, P. (2020). *Conserving Africa’s wildlife and wildlands*

599 through the COVID-19 crisis and beyond. *Nature Ecology & Evolution*, 4(10), 10.
600 <https://doi.org/10.1038/s41559-020-1275-6>

601 Loftus, J.C., Harel, R., Núñez, C.L., Crofoot, M.C., 2022. Ecological and social pressures interfere with
602 homeostatic sleep regulation in the wild. *eLife* 11, e73695.

603 Łopucki, R., Kitowski, I., Perlińska-Teresiak, M., & Klich, D. (2021). How Is Wildlife Affected by the
604 COVID-19 Pandemic? Lockdown Effect on the Road Mortality of Hedgehogs. *Animals*, 11(3), 3.
605 <https://doi.org/10.3390/ani11030868>

606 Lüdecke, D. (2018). ggeffects: Tidy Data Frames of Marginal Effects from Regression Models. *Journal*
607 *of Open Source Software*, 3(26), 772. <https://doi.org/10.21105/joss.00772>

608 Lutermann, H., Verburgt, L., & Rendigs, A. (2010). Resting and nesting in a small mammal: Sleeping sites
609 as a limiting resource for female grey mouse lemurs. *Animal Behaviour*, 79(6), 1211–1219.
610 <https://doi.org/10.1016/j.anbehav.2010.02.017>

611 Manenti, R., Mori, E., Di Canio, V., Mercurio, S., Picone, M., Caffi, M., Brambilla, M., Ficetola, G. F.,
612 & Rubolini, D. (2020). The good, the bad and the ugly of COVID-19 lockdown effects on wildlife
613 conservation: Insights from the first European locked down country. *Biological Conservation*,
614 249, 108728. <https://doi.org/10.1016/j.biocon.2020.108728>

615 Martin, S.E., Wraith, P.K., Deary, I.J., Douglas, N.J., 2012. The effect of nonvisible sleep fragmentation
616 on daytime function. *American journal of respiratory and critical care medicine*.

617 Massei, G., Kindberg, J., Licoppe, A., Gačić, D., Šprem, N., Kamler, J., Baubet, E., Hohmann, U., Monaco,
618 A., Ozoliņš, J., Cellina, S., Podgórski, T., Fonseca, C., Markov, N., Pokorny, B., Rosell, C., & Náhlik,
619 A. (2015). Wild boar populations up, numbers of hunters down? A review of trends and
620 implications for Europe. *Pest Management Science*, 71(4), 492–500.
621 <https://doi.org/10.1002/ps.3965>

622 Max-Planck-Gesellschaft. (2021, September 8). *Tierische Pause vom Menschen*.
623 <https://www.mpg.de/15005711/covid-19-bio-logging-initiative>

624 McDermott, C. M., LaHoste, G. J., Chen, C., Musto, A., Bazan, N. G., & Magee, J. C. (2003). Sleep
625 deprivation causes behavioral, synaptic, and membrane excitability alterations in hippocampal
626 neurons. *The Journal of Neuroscience: The Official Journal of the Society for Neuroscience*,
627 23(29), 9687–9695.

628 McGinlay, J., Gkoumas, V., Holtvoeth, J., Fuertes, R. F. A., Bazhenova, E., Benzoni, A., Botsch, K., Martel,
629 C. C., Sánchez, C. C., Cervera, I., Chaminade, G., Doerstel, J., García, C. J. F., Jones, A., Lammertz,
630 M., Lotman, K., Odar, M., Pastor, T., Ritchie, C., ... Jones, N. (2020). The Impact of COVID-19 on
631 the Management of European Protected Areas and Policy Implications. *Forests*, 11(11), 11.
632 <https://doi.org/10.3390/f11111214>

633 Miwa, M., Oishi, K., Anzai, H., Kumagai, H., Ieiri, S., & Hirooka, H. (2017). Estimation of the energy
634 expenditure of grazing ruminants by incorporating dynamic body acceleration into a
635 conventional energy requirement system1. *Journal of Animal Science*, 95(2), 901–909.
636 <https://doi.org/10.2527/jas.2016.0749>

637 Morelle, K., Fattebert, J., Mengal, C., & Lejeune, P. (2016). Invading or recolonizing? Patterns and
638 drivers of wild boar population expansion into Belgian agroecosystems. *Agriculture,*
639 *Ecosystems & Environment*, 222, 267–275. <https://doi.org/10.1016/j.agee.2016.02.016>

640 Mortlock, E., Silovský, V., Güldenpfennig J., Faltusová M., Olejarz A., Börger L., Ježek M., Jennings D.,
641 Capellini I. (2022). Individual identity and environmental conditions explain different aspects
642 of sleep behaviour in wild boar. Preprint at <https://doi.org/10.1101/2022.11.23.517569>

643 Phillipson, E.A., Bowes, G., Sullivan, C.E., Woolf, G.M., 1980. The Influence of Sleep Fragmentation on
644 Arousal and Ventilatory Responses to Respiratory Stimuli. *Sleep* 3, 281–288.

645 Podgórski, T., Baś, G., Jędrzejewska, B., Sönnichsen, L., Śnieżko, S., Jędrzejewski, W., & Okarma, H.
646 (2013). Spatiotemporal behavioral plasticity of wild boar (*Sus scrofa*) under contrasting
647 conditions of human pressure: Primeval Forest and metropolitan area. *Journal of Mammalogy*,
648 94(1), 109–119. <https://doi.org/10.1644/12-MAMM-A-038.1>

649 Podrázský, V., Remeš, J., Hart, V., & Moser, W. (2009). Production and humus form development in
650 forest stands established on agricultural lands—Kostelec nad Černými lesy region. *Journal of*
651 *Forest Science*, 55, 299–305. <https://doi.org/10.17221/11/2009-JFS>

652 Qasem, L., Cardew, A., Wilson, A., Griffiths, I., Halsey, L. G., Shepard, E. L. C., Gleiss, A. C., & Wilson, R.
653 (2012). Tri-Axial Dynamic Acceleration as a Proxy for Animal Energy Expenditure; Should We
654 Be Summing Values or Calculating the Vector? *PLoS ONE*, 7(2).
655 <https://doi.org/10.1371/journal.pone.0031187>

656 R Core Team. (2021). *R: A language and environment for statistical computing*. R Foundation for
657 *Statistical Computing*.

658 Rahman, Md. S., Alam, Md. A., Salekin, S., Belal, Md. A. H., & Rahman, Md. S. (2021). The COVID-19
659 pandemic: A threat to forest and wildlife conservation in Bangladesh? *Trees, Forests and*
660 *People*, 5, 100119. <https://doi.org/10.1016/j.tfp.2021.100119>

661 Rechtschaffen, A., Bergmann, B.M., 2002. Sleep Deprivation in the Rat: An Update of the 1989 Paper.
662 *Sleep* 25, 18–24.

663 Rogers, N. L., Szuba, M. P., Staab, J. P., Evans, D. L., & Dinges, D. F. (2001). Neuroimmunologic aspects
664 of sleep and sleep loss. *Seminars in Clinical Neuropsychiatry*, 6(4), 295–307.
665 <https://doi.org/10.1053/scnp.2001.27907>

666 Rosell, C., Navàs, F., & Romero, S. (2012). Reproduction of wild boar in a cropland and coastal wetland
667 area: Implications for management. *Animal Biodiversity and Conservation*, 35, 209–217.
668 <https://doi.org/10.32800/abc.2012.35.0209>

669 Roth, T. C., Rattenborg, N. C., & Pravosudov, V. V. (2010). The ecological relevance of sleep: The trade-
670 off between sleep, memory and energy conservation. *Philosophical Transactions of the Royal*
671 *Society B: Biological Sciences*, 365(1542), 945–959. <https://doi.org/10.1098/rstb.2009.0209>

672 Russo, L., Massei, G., & Genov, P. (2010). Daily home range and activity of wild boar in a Mediterranean
673 area free from hunting. *Ethology Ecology & Evolution, July 1997*, 287–294.
674 <https://doi.org/10.1080/08927014.1997.9522888>

675 Rutz, C. (2022). Studying pauses and pulses in human mobility and their environmental impacts. *Nature*
676 *Reviews Earth & Environment*, 3, 157–159. <https://doi.org/10.1038/s43017-022-00276-x>

677 Saïd, S., Tolon, V., Brandt, S., Baubet, E., 2012. Sex effect on habitat selection in response to hunting
678 disturbance: the study of wild boar. *Eur J Wildl Res* 58, 107–115.

679 Said, M. Y., Ogutu, J. O., Kifugo, S. C., Makui, O., Reid, R. S., & de Leeuw, J. (2016). Effects of extreme
680 land fragmentation on wildlife and livestock population abundance and distribution. *Journal*
681 *for Nature Conservation*, 34, 151–164. <https://doi.org/10.1016/j.jnc.2016.10.005>

682 Scandura, Tomasz Podgorski, Joaquin Vicente, & Laura Iacolina. (2021). *Wild Boar Sus scrofa Linnaeus,*
683 *1758. Handbook of the Mammals of Europe K. Hackländer and F. E. Zacos*. Cham, Springer
684 International Publishing: 1-28. <https://link.springer.com/book/10.1007/978-3-319-65038-8>

685 Scholten, J., Moe, S., & Hegland, S. (2018). Red deer (*Cervus elaphus*) avoid mountain biking trails.
686 *European Journal of Wildlife Research*, 64. <https://doi.org/10.1007/s10344-018-1169-y>

687 Scillitani, L., Monaco, A., & Toso, S. (2009). Do intensive drive hunts affect wild boar () spatial behaviour
688 in Italy? Some evidences and management implications. *European Journal of Wildlife Research*,
689 56(3), 307–318. <https://doi.org/10.1007/s10344-009-0314-z>

690 Seip, D. R., Johnson, C. J., & Watts, G. S. (2007). Displacement of Mountain Caribou From Winter
691 Habitat by Snowmobiles. *The Journal of Wildlife Management*, 71(5), 1539–1544.
692 <https://doi.org/10.2193/2006-387>

693 Shi, H., Shi, T., Yang, Z., Wang, Z., Han, F., & Wang, C. (2018). Effect of Roads on Ecological Corridors
694 Used for Wildlife Movement in a Natural Heritage Site. *Sustainability*, 10(8), 8.
695 <https://doi.org/10.3390/su10082725>

696 Sibbald, A. M., Hooper, R. J., McLeod, J. E., & Gordon, I. J. (2011). Responses of red deer (*Cervus*
697 *elaphus*) to regular disturbance by hill walkers. *European Journal of Wildlife Research*, 57(4),
698 817–825. <https://doi.org/10.1007/s10344-011-0493-2>

699 Singer, F. J., Otto, D. K., Tipton, A. R., & Hable, C. P. (1981). Home Ranges, Movements, and Habitat Use
700 of European Wild Boar in Tennessee. *The Journal of Wildlife Management*, 45(2), 343–353.
701 <https://doi.org/10.2307/3807917>

702 Sodeikat, G., Pohlmeier, K., 2002. Temporary home range modifications of wild boar family groups
703 (*Sus scrofa* L.) caused by drive hunts in Lower Saxony (Germany). *Zeitschrift für*
704 *Jagdwissenschaft* 48, 161–166

705 Sodeikat, G., & Pohlmeier, K. (2003). Escape movements of family groups of wild boar *Sus scrofa*
706 influenced by drive hunts in Lower Saxony, Germany. *Wildlife Biology*, 9(4), 43–49.
707 <https://doi.org/10.2981/wlb.2003.063>

708 Sodeikat, G., & Pohlmeier, K. (2007). Impact of drive hunts on daytime resting site areas of wild boar
709 family groups (*Sus scrofa* L.). *Wildlife Biology in Practice; Vol 3, No 1 (2007); 28-38, 3.*
710 <https://doi.org/10.2461/wbp.2007.3.4>

711 Stillfried, M., Gras, P., Busch, M., Börner, K., Kramer-Schadt, S., Ortman, S., 2017. Wild inside: Urban
712 wild boar select natural, not anthropogenic food resources. *PLOS ONE* 12, e0175127.

713 Thurfjell, H., Spong, G., & Ericsson, G. (2013). Effects of hunting on wild boar *Sus scrofa* behaviour.
714 *Wildlife Biology*, 19(1), 87–93. <https://doi.org/10.2981/12-027>

715 Thurfjell, H., Spong, G., & Ericsson, G. (2014). Effects of weather, season, and daylight on female wild
716 boar movement. *Acta Theriologica*, 59(3), 467–472. <https://doi.org/10.1007/s13364-014-0185-x>

717

718 Tucker, M. A., Böhning-Gaese, K., Fagan, W. F., Fryxell, J. M., Van Moorter, B., Alberts, S. C., Ali, A. H.,
719 Allen, A. M., Attias, N., Avgar, T., Bartlam-Brooks, H., Bayarbaatar, B., Belant, J. L., Bertassoni,
720 A., Beyer, D., Bidner, L., van Beest, F. M., Blake, S., Blaum, N., ... Mueller, T. (2018). Moving in
721 the Anthropocene: Global reductions in terrestrial mammalian movements. *Science*,
722 359(6374), 466–469. <https://doi.org/10.1126/science.aam9712>

723 Tuomainen, U., Candolin, U., 2011. Behavioural responses to human-induced environmental change.
724 *Biological Reviews* 86, 640–657

725 Venter, Z. S., Barton, D. N., Gundersen, V., Figari, H., & Nowell, M. (2020). Urban nature in a time of
726 crisis: Recreational use of green space increases during the COVID-19 outbreak in Oslo,
727 Norway. *Environmental Research Letters*, 15(10), 104075. <https://doi.org/10.1088/1748-9326/abb396>

728

729 Vitousek, P. M., Mooney, H. A., Lubchenco, J., & Melillo, J. M. (1997). Human Domination of Earth's
730 Ecosystems. *Science*, 277(5325), 494–499. <https://doi.org/10.1126/science.277.5325.494>

731 vlada.cz. (2020). *Measures adopted by the Czech Government against the coronavirus | Government*
732 *of the Czech Republic.* [https://www.vlada.cz/en/media-centrum/aktualne/measures-adopted-](https://www.vlada.cz/en/media-centrum/aktualne/measures-adopted-by-the-czech-government-against-coronavirus-180545#general)
733 [by-the-czech-government-against-coronavirus-180545#general](https://www.vlada.cz/en/media-centrum/aktualne/measures-adopted-by-the-czech-government-against-coronavirus-180545#general)

734 Voirin, B., Scriba, M.F., Martinez-Gonzalez, D., Vyssotski, A.L., Wikelski, M., Rattenborg, N.C., 2014.
735 Ecology and Neurophysiology of Sleep in Two Wild Sloth Species. *Sleep* 37, 753–761.

736 Weed, M. (2020). The role of the interface of sport and tourism in the response to the COVID-19
737 pandemic. *Journal of Sport & Tourism*, 24(2), 79–92.
738 <https://doi.org/10.1080/14775085.2020.1794351>

739 Wildbyte Technologies. (2022). *Software.* <http://www.wildbytetechnologies.com/software.html>

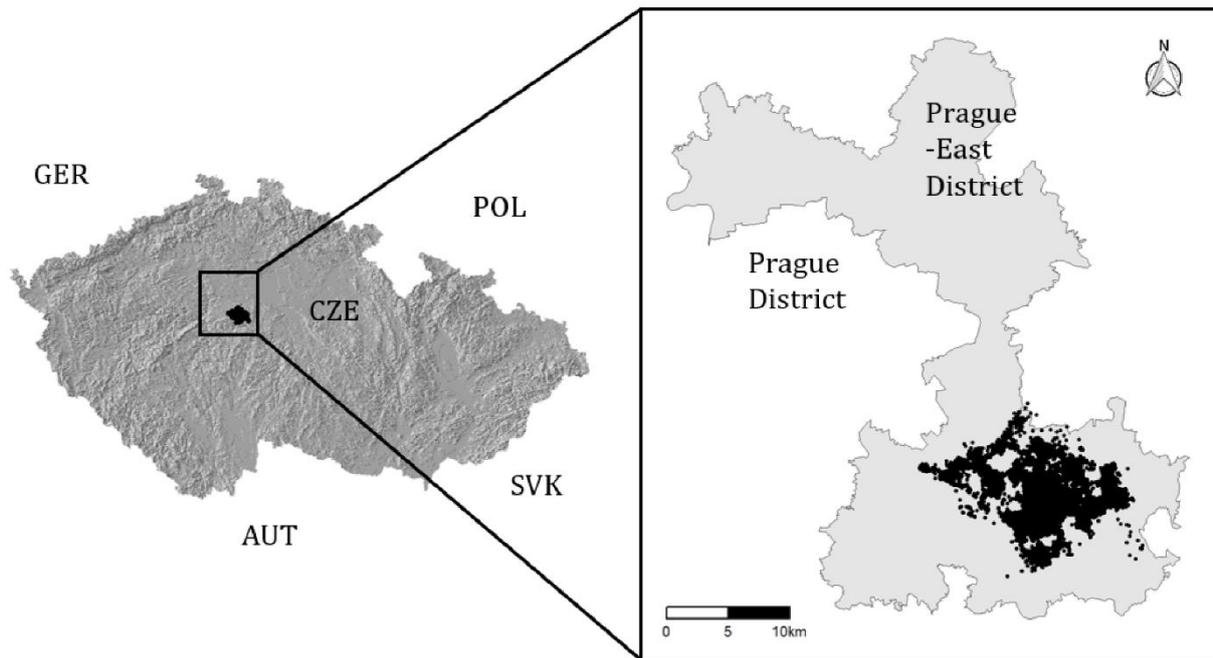
740 Wilson, R.P., Börger, L., Holton, M.D., Scantlebury, D.M., Gómez-Laich, A., Quintana, F., Rosell, F., Graf,
741 P.M., Williams, H., Gunner, R., Hopkins, L., Marks, N., Gerdali, N.R., Duarte, C.M., Scott, R.,
742 Strano, M.S., Robotka, H., Eizaguirre, C., Fahlman, A., Shepard, E.L.C., 2020. Estimates for
743 energy expenditure in free-living animals using acceleration proxies: A reappraisal. *Journal of*
744 *Animal Ecology* 89, 161–172.

745 Zaid, E., Vyssotski, A. L., & Lesku, J. A. (2022). Sleep architecture and regulation of male dusky
746 antechinus, an Australian marsupial. *Sleep*, 45(8), zsac114.
747 <https://doi.org/10.1093/sleep/zsac114>

748 Zukerman, Y., Sigal, Z., & Berger-Tal, O. (2021). COVID-19 Restrictions in a Nature Reserve Reveal the
749 Costs of Human Presence for the Threatened Nubian Ibex (*Capra nubiana*). *Frontiers in Ecology*
750 *and Evolution*, 9. <https://www.frontiersin.org/articles/10.3389/fevo.2021.751515>

751

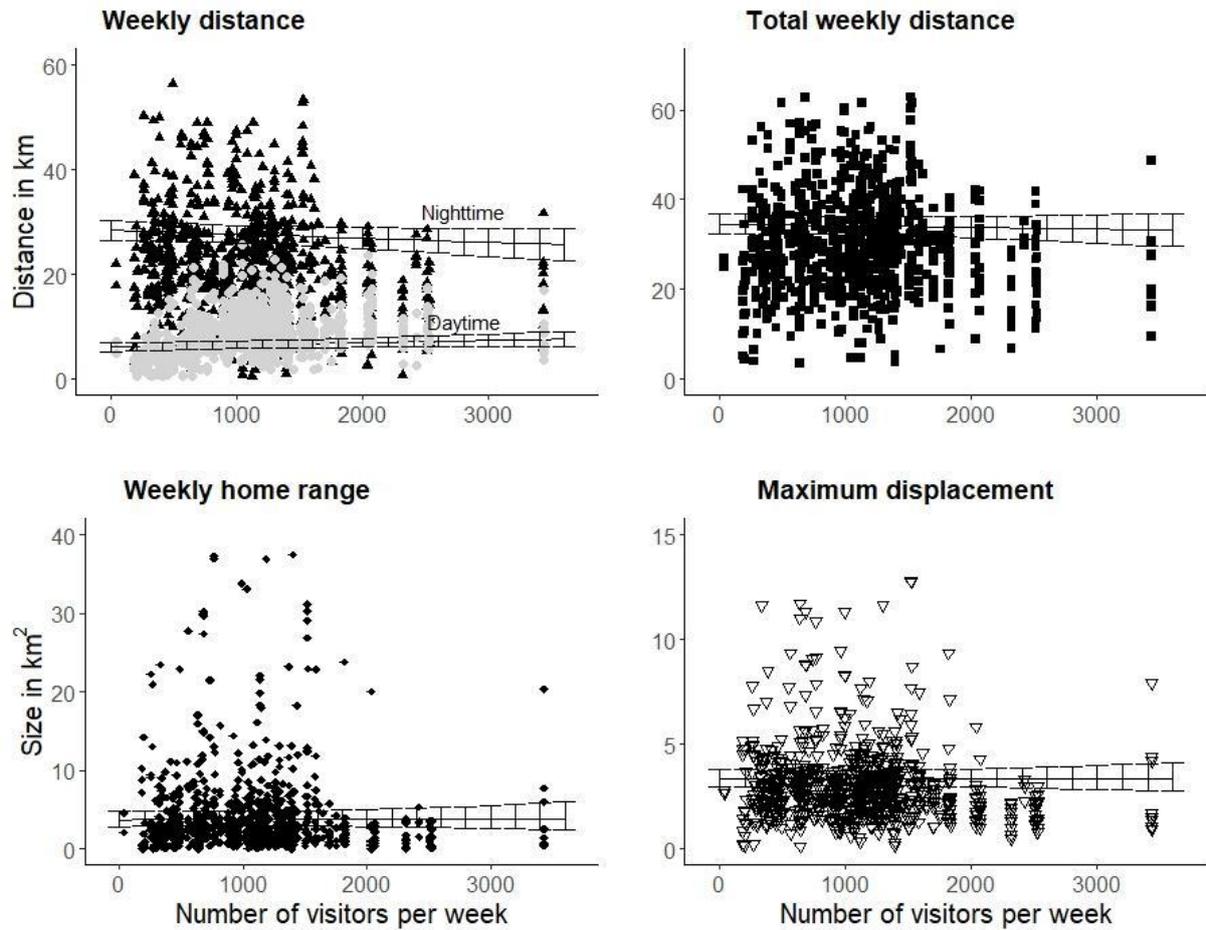
752 **Appendices**



753

754 **Figure A.1: Location of the study area Prague-East in the Czech Republic (CZE) with**
755 **GPS positions of the collared wild boars (black points).**

756



757

758 **Figure A.2: Changes in the movement of wild boar in relation to numbers of visitors per**
 759 **week: A) total weekly distance at nighttime and daytime B) total weekly distance C)**
 760 **weekly home range 95 % Kernel D) maximum displacement (maximum distance of GPS**
 761 **locations within a week)**