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1 Machine learning-based rapid response tools for regional
2 air pollution modelling

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16 **Abstract**

17 A parameterised non-intrusive reduced order model (P-NIROM) based on
18 proper orthogonal decomposition (POD) and machine learning methods has
19 been firstly developed for model reduction of pollutant transport equations.
20 Our motivation is to provide rapid response urban air pollution predictions
21 and controls. The varying parameters in the P-NIROM are pollutant sources.
22 The training data sets are obtained from the high fidelity modelling solutions
23 (called snapshots) for selected parameters (pollutant sources, here) over the
24 parameter space \mathcal{R}^P . From these training data sets, the machine learning
25 method is used to generate the relationship between the reduced solutions
26 and inputs (pollutant sources) over \mathcal{R}^P . Furthermore a set of hyper-
27 surface functions associated with each POD basis function is constructed for
28 representing the fluid dynamics over the reduced space. The accuracy of
29 the P-NIROM is highly dependent on the quality of the training set, here
30 obtained from the high fidelity model. Over existing machine learning meth-
31 ods, the P-NIROM algorithm proposed here has the advantages that (1) it is
32 combined with NIROM, thus providing rapid and reasonably accurate solu-
33 tions; and (2) it is a robust and efficient approach for representation of any
34 parametrised partial differential equations as the model parameters/inputs

35 vary. In this study, we demonstrate the way how to implement the P-NIROM
36 for the pollutant transport equation (but not limited to due to its robust-
37 ness). Its predictive capability is illustrated in a three-dimensional (3-D)
38 simulation of power plant plumes over a large region in China, where the
39 varying parameters are the emission intensity at three locations. Results in-
40 dicate that in comparison to the high fidelity model, the CPU cost is reduced
41 by factor up to five orders of magnitude while reasonable accuracy remains.

42 *Keywords:* Machine learning, Finite Element, Proper orthogonal
43 decomposition, Reduced order modelling, air pollution.

44 1. Introduction

45 Pollution in cities has a strong impact on the health of communities and
46 affects global warming with dire consequences to humanity. The dynamic
47 and pollutant transport processes involve a wide range of scales. The highly
48 disparate scale poses a formidable challenge for atmospheric and air pollution
49 modelling. In recent years, the spatial resolution in operation air pollution
50 models has been increased significantly, thus improving predictive capability.
51 However, this unavoidably leads to an increase in computational cost [1].
52 Our motivation is to develop numerical tools for rapid responses/predictions
53 of pollutants without sacrificing solution accuracy, especially in emergency
54 situations.

55 Reduced-order models (ROMs) have become important to many fields
56 as they offer the potential to simulate dynamical systems with considerably
57 reduced computational cost in comparison to high fidelity models [2, 3, 4, 5].
58 Recently, reduced order methods have been applied to studies of air pollution
59 [6, 7, 8, 9]. Existing ROMs can be classified into two categories: intrusive
60 and non-intrusive approaches in the sense that whether the implementation
61 of ROMs requires knowledge of the details of original numerical source codes
62 [10]. The intrusive reduced order methods have been widely used in many
63 fields [11, 12, 13, 14, 15, 16, 17, 18]. More recently, the non-intrusive methods
64 have become popular since they are less dependent on complex dynamic
65 systems and are therefore easy to implement even when the numerical source
66 code is not available. Existing non-intrusive methods used for generating
67 ROMs are POD in combination with radial basis function (RBF), Smolyak
68 sparse grid and artificial neural networks etc.[19, 20, 21, 22, 23, 24]. The
69 applications of NIROMs can be found in the work of [25, 26, 27, 28].

70 More recently, Wang *et al.* introduced a deep learning technique to
71 NIROMs and applied it to fluid problems [29]. Deep learning technologies
72 represent the most recent progress in artificial neural networks [30], and have
73 been applied to a number of areas such as speech recognition [31], image
74 recognition [32], medical science [33], self-driving cars [34], language under-
75 standing [35] etc.

76 In this work, we have developed a Parameterised NIROM (P-NIROM)
77 based on machine learning techniques for parameterised pollutant transport
78 problems. The input parameters are the emission intensity of pollutants re-
79 leased at different source locations. The P-NIROM enables rapid simulations
80 and controls of the impact of pollutant sources without excessive computa-
81 tional costs. Given a set of selected pollutant sources \mathbf{Q}_{tr} over the parame-
82 terised space R^P , the training data sets (also called solution snapshots) can
83 be obtained by running the high fidelity model. From the snapshot solutions,
84 the corresponding reduced basis functions are calculated using singular value
85 decomposition (SVD)/POD. The reduced basis functions are used for con-
86 structing the reduced space. The original high fidelity model can be projected
87 onto the reduced space, which is several orders of magnitude smaller than the
88 dimensional size of the high fidelity full model, thus significantly reducing the
89 computational cost. For any unseen emission intensity of pollutant sources
90 $\mathbf{Q} \in \mathcal{R}^P$, the P-NIROM is constructed using the machine learning methods.
91 From the training solution snapshots, a Gaussian process is used for generat-
92 ing the snapshots and POD basis functions for the unseen pollutant sources
93 \mathbf{Q} . Furthermore, the relationship (P-NIROM) between the reduced solutions
94 and the inputs (the pollutant emission intensities) can be obtained using the
95 machine learning techniques. Finally, the solutions from the P-NIROM are
96 projected back the full space.

97 The P-NIROM is a robust and efficient numerical tool for rapid prediction
98 of pollutants released from different sources and assessment of their impact on
99 specified cities/locations. In this work, we have been successfully applied the
100 P-NIROM to air pollution simulations over a large region in China which
101 covers 55 cities including Beijing. The efficiency and accuracy of the P-
102 NIROM have been evaluated by comparing the results with those from the
103 high fidelity full model.

104 The remainder of this article is arranged as follows. The pollutant trans-
105 port equation and its discretisation are described in section 2. In section 3,
106 the details of forming the P-NIROM using POD and machine learning meth-
107 ods are provided. Section 4 presents a numerical experiment of simulating

108 the spatial and temporary distribution of pollutants released from 100 power
 109 plants in China. Conclusions are drawn in section 5.

110 2. Pollutant transport equation and its discretisation

The dispersion of the tracer concentration (c) is modelled by:

$$\frac{\partial c}{\partial t} + \mathbf{u} \cdot \nabla c + \nabla \cdot (\kappa \nabla c) - Q = 0, \quad (1)$$

where \mathbf{u} is the velocity vector, Q is a source term and κ the diffusivity. In general, the discretised form of (1) at each time level n (where a time interval of Δt is set during the simulation period) can be written:

$$\mathcal{M}(\mu) \mathbf{c}^n = \mathbf{s}^n(\mu, \mathbf{c}^{n-1}), \quad (2)$$

111 where \mathcal{M} is the full numerical operator with varying input parameter μ ,
 112 $\mathbf{c}^n = (c_1^n \dots, c_j^n, \dots, c_{\mathcal{N}}^n)^T$ ($1 \leq j \leq \mathcal{N}$, \mathcal{N} is the number of nodes in the
 113 computational domain), \mathbf{s}^s includes the source term, boundary conditions
 114 and the variable solutions from the previous time level. In this study, the
 115 varying input parameters in air pollutant problems are set to be the pollutant
 116 sources, $\mu = \mathbf{Q} = (Q_1, \dots, Q_s, \dots, Q_S)$ (here, S is the number of pollutant
 117 sources).

118 2.1. Parameterised reduced order transport equation

In this work, the POD approach in combination with machine learning techniques is used for model reduction. POD has proven to be a powerful tool for circumventing the intensive computational burden in large complex numerical simulations. POD is capable of representing large complex dynamical systems using a few number of optimal basis functions. In POD reduced order modelling, the tracer concentration in (2) can be expressed as an expansion of the POD basis functions $\Phi = (\Phi_1, \dots, \Phi_m, \dots, \Phi_M)$:

$$\mathbf{c}^n = \Phi \mathbf{c}^{r,n}, \quad (3)$$

where $\mathbf{c}^{r,n} = (c_1^{r,n} \dots, c_m^{r,n}, \dots, c_M^{r,n})^T$ ($1 \leq m \leq M$) $\in \mathcal{R}^M$ is the reduced state variable vector (the superscript r indicates the variable associated with the reduced order model) to be determined over the reduced space. The POD basis functions are constructed from a collection of snapshots that are taken from the high fidelity model solution (2) for the selected training pollutant

sources. Using SVD, a set of orthogonal basis functions $\{\Phi_m\}$ can be obtained in an optimal way. The POD basis functions can represent the dynamics of snapshot solutions. The loss of information due to the truncation of the POD expansion set to M vectors can be quantified by the following ratio,

$$E = \frac{\sum_{j=1}^M \lambda_j^2}{\sum_{j=1}^I \lambda_j^2}, \quad (4)$$

119 where λ denotes eigenvalues, and I is the total number of eigenvectors (here
 120 equivalent to the number of solution snapshots used for generating the POD
 121 basis functions). The value of E will tend to 1 as M is increased to the value I ,
 122 this would imply no loss of information. A few number of leading eigenvectors
 123 can represent most of dynamical energy within the solution snapshots.

Projecting (2) from the \mathcal{N} dimensional space onto the M dimensional reduced space ($M \ll \mathcal{N}$), yields:

$$\Phi^T \mathcal{M}(\mu) \mathbf{c}^n = \Phi^T \mathbf{s}^n(\mu, \mathbf{c}^{n-1}). \quad (5)$$

The parameterised reduced order model can thus be written as:

$$\mathcal{M}^r(\mu) \mathbf{c}^{r,n} = \mathbf{s}^{r,n}(\mu, \mathbf{c}^{r,n-1}), \quad (6)$$

124 where $\mathcal{M}_{t_0, t_j}^r = \Phi^T \mathcal{M}(\mu)$ is the model operator over the reduced space, $\mathbf{c}^{r,n} =$
 125 $\Phi^T \mathbf{c}^n$ and $\mathbf{s}^{r,n} = \Phi^T \mathbf{s}^n(\mu, \mathbf{c}^{r,n-1})$.

126 Equations (3) and (6) can be used for efficient air pollution operational
 127 prediction where the CPU time can be reduced by several orders of magni-
 128 tude. In this work, the parameter set μ in (6) consists of the pollutant source
 129 inputs. A recently developed NIROM [29] is extended to construct the pa-
 130 rameterised reduced order model in (6). The P-NIROM based on the machine
 131 learning techniques described below is capable of predicting problems with
 132 unseen or different parameters (for example, unseen pollutant sources). It is
 133 also non-intrusive and independent of the original source code.

134 3. Construction of P-NIROM based on POD and machine learning 135 methods

The parameterised reduced order model (6) is re-written for the variable c_m^r associated with each POD basis function Φ_m over the reduced space in a general form:

$$c_m^{r,n} = \mathcal{F}_m(\mu, \mathbf{c}^{r,n-1}), \quad m \in (1, \dots, M). \quad (7)$$

136 In non-intrusive reduced order modelling, one searches a set of functions
 137 \mathcal{F}_m to represent the dynamics in (7). In this work, we introduce the Gaussian
 138 process regression (GPR) [36] and deep learning learning methods [29] to
 139 construct the relationship functions \mathcal{F}_m to represent the fluid dynamics of
 140 system (6) for any unseen input parameter $\mu = \mathbf{Q} \in \mathcal{R}^P$.

141 *3.1. Gaussian process regression for calculation of POD coefficient and snap-*
 142 *shot solutions for any input over the parameter space*

In GPR, the relationship between the input $\mu = \mathbf{Q}$ (here, the pollutant source) and the corresponding output \mathbf{c}^n at each time level $n \in (1, 2, \dots, T_{tr})$ can be expressed as follows [36]:

$$\mathbf{c}^n(\mathbf{Q}) = g^n(\mathbf{Q}) + \epsilon(\mathbf{Q}), \quad (8)$$

143 where, $\epsilon = \mathcal{G}(0, \sigma_n)$ is the Gaussian distribution with zero mean and variance
 144 σ_n .

In GPR, it is assumed that the function $g^n(\mathbf{Q})$ has a Gaussian distribution (with zero mean, here):

$$g^n(\mathbf{Q}) \sim \mathcal{G}(\mathbf{0}, k^n(\mathbf{Q}, \mathbf{Q}')), \quad (9)$$

where the covariance function $k^n(\mathbf{Q}, \mathbf{Q}')$ represents the dependency between the function values at two different input points \mathbf{Q} and \mathbf{Q}' , that is,

$$cov(g^n(\mathbf{Q}), g^n(\mathbf{Q}')) = k^n(\mathbf{Q}, \mathbf{Q}') = \sigma_w^n \exp\left(-\frac{1}{2l}|\mathbf{Q} - \mathbf{Q}'|\right), \quad (10)$$

where, l is the length scale and σ_w^n is the variance. The correlation between the functions $g^n(\mathbf{Q})$ and $g^n(\mathbf{Q}')$ is dependent on the distance between the two input points. Given a set of training input-output pairs $\{\mathbf{Q}_{tr,i}, \mathbf{c}_{tr,i}^n\}, i \in (1, \dots, N_{tr})$ (where, N_{tr} is the number of training points), one aims to predict the pollutant concentration \mathbf{c}^n in (8) for any new input \mathbf{Q} . The joint Gaussian distribution of the training and predicted outputs (\mathbf{c}_{tr}^n and \mathbf{c}^n) for the training and new inputs (\mathbf{Q}_{tr} and \mathbf{Q}) respectively can be written:

$$\begin{bmatrix} \mathbf{c}_{tr}^n \\ \mathbf{c}^n \end{bmatrix} = \begin{bmatrix} g^n(\mathbf{Q}_{tr}) \\ g^n(\mathbf{Q}) \end{bmatrix} \sim \mathcal{N}\left(\mathbf{0} \begin{bmatrix} K^n(\mathbf{Q}_{tr}, \mathbf{Q}_{tr}) & K^{nT}(\mathbf{Q}, \mathbf{Q}_{tr}) \\ K^n(\mathbf{Q}, \mathbf{Q}_{tr}) & K^n(\mathbf{Q}, \mathbf{Q}) \end{bmatrix}\right), \quad (11)$$

where, $K^n(\mathbf{Q}_{tr}, \mathbf{Q}_{tr})$ is the covariance matrix between all training points and is written below:

$$K^n(\mathbf{Q}_{tr}, \mathbf{Q}_{tr}) = \begin{bmatrix} k^n(\mathbf{Q}_{tr,1}, \mathbf{Q}_{tr,1}) & k^n(\mathbf{Q}_{tr,1}, \mathbf{Q}_{tr,2}) & \dots & k^n(\mathbf{Q}_{tr,1}, \mathbf{Q}_{tr,N_{tr}}) \\ k^n(\mathbf{Q}_{tr,2}, \mathbf{Q}_{tr,1}) & k^n(\mathbf{Q}_{tr,2}, \mathbf{Q}_{tr,2}) & \dots & k^n(\mathbf{Q}_{tr,2}, \mathbf{Q}_{tr,N_{tr}}) \\ \vdots & \vdots & \ddots & \vdots \\ k^n(\mathbf{Q}_{tr,N_{tr}}, \mathbf{Q}_{tr,1}) & k^n(\mathbf{Q}_{tr,N_{tr}}, \mathbf{Q}_{tr,2}) & \dots & k^n(\mathbf{Q}_{tr,N_{tr}}, \mathbf{Q}_{tr,N_{tr}}) \end{bmatrix}, \quad (12)$$

and the matrices

$$K^n(\mathbf{Q}_{tr}, \mathbf{Q}) = [k^n(\mathbf{Q}_{tr,1}, \mathbf{Q}) \quad k^n(\mathbf{Q}_{tr,2}, \mathbf{Q}) \quad \dots \quad k^n(\mathbf{Q}_{tr,N_{tr}}, \mathbf{Q})], \quad (13)$$

$$K^n(\mathbf{Q}, \mathbf{Q}) = k^n(\mathbf{Q}, \mathbf{Q}). \quad (14)$$

145 Given a set of the training inputs (here, the pollutant sources) $\mu_{tr} =$
 146 $\mathbf{Q}_{tr} = (Q_{tr,1}, \dots, Q_{tr,S})$ over the parameter space R^P , the snapshot solutions
 147 $\mathbf{c}_{tr} = (\mathbf{c}_{tr}^1, \dots, \mathbf{c}_{tr}^n, \dots, \mathbf{c}_{tr}^{N_t})$ can be obtained by running the high fidelity model
 148 (2) during the training simulation period $[0, T_{tr}]$.

For efficient calculations, one can project \mathbf{c}_{tr}^n from the high dimensional full space onto the reduced space:

$$\mathbf{c}_{tr}^{r,n} = \Phi^T \mathbf{c}_{tr}^n, \quad (15)$$

For any given input parameter (pollution source \mathbf{Q}), the probability of the prediction of the reduced variable \mathbf{c}^r is:

$$\mathbf{c}_{tr}^{r,n} | \mathbf{c}_{tr}^{r,n} \sim \mathcal{N}(K_*^n K_{tr}^{n-1} \mathbf{c}_{tr}^{r,n}, K_{**}^n - K_*^n K_{tr}^{n-1} K_*^{nT}), \quad (16)$$

where, $K_*^n = K^n(\mathbf{Q}, \mathbf{Q}_{tr})$, $K_{**}^n = K^n(\mathbf{Q}, \mathbf{Q})$ and $K_{tr}^n = K^n(\mathbf{Q}_{tr}, \mathbf{Q}_{tr})$. The best estimate of $\mathbf{c}^{r,n}$ is the mean of the Gaussian distribution:

$$\bar{\mathbf{c}}^{r,n} = K_*^n K_{tr}^{n-1} \mathbf{c}_{tr}^{n,r}. \quad (17)$$

149 3.2. Deep learning method for construction of P-NIROM and calculation of 150 reduced solutions for any input over the parameter space

151 In this section, an alternative method for calculation of reduced solutions
 152 for any given input is introduced. A Recurrent Neural Network (RNN) using
 153 the Long Short Term Memory (LSTM) architecture is used to construct the
 154 P-NIROM (7). Compared to traditional RNNs, the LSTM has a special

155 memory block in the hidden layer of the recurrent neural network, allowing
 156 information to persist. This type of network has cyclic connections, which
 157 makes the network a powerful method to model temporal data since it has an
 158 internal memory system to deal with temporal sequence inputs. A memory
 159 cell is composed of four main elements: an input gate, a neuron with a self-
 160 recurrent connection (a connection to itself), a forget gate and an output
 161 gate.

162 The input gate of each memory block controls the information transmit-
 163 ting from the input activations into the memory cell and the output gate
 164 controls the information transmitting from the memory cell activations into
 165 other nodes. The forget gate decides what information is to be deleted from
 166 the memory cell state [29].

167 The LSTM technique is utilised to construct the set of functions (hyper-
 168 surfaces) F_m in (7). In the LSTM network, the input is the reduced solution
 169 $\mathbf{c}^{r,n-1} = (c_1^{r,n-1}, \dots, c_M^{r,n-1})$ at the previous time level $n - 1$ while the output
 170 is the reduced solution $c_m^{r,n}$ associated with the m^{th} POD basis function Φ_m
 171 ($m \in (1, \dots, M)$). The relationship function (hyper-surface F_m) between the
 172 input $\mathbf{c}^{r,n-1}$ and output $c_m^{r,n}$ can be obtained using the following equations:

$$\begin{aligned}
 I^n &= \varrho(W_{ic}\mathbf{c}^{r,n-1} + W_{ih}h^{n-1} + W_{iCe}Ce^{n-1} + b_i), \\
 f^n &= \varrho(W_{fc}\mathbf{c}^{r,n-1} + W_{fh}h^{n-1} + W_{fCe}Ce^{n-1} + b_r), \\
 o^n &= \varrho(W_{oc}\mathbf{c}^{r,n-1} + W_{oh}h^{n-1} + W_{oCe}Ce^{n-1} + b_o), \\
 Ce^n &= r^n \odot Ce^{n-1} + i^n \odot Ce_i(W_{cec}\mathbf{c}^{r,n-1} + W_{ceh}h^{n-1} + b_{Ce}), \\
 h^n &= o^n \odot Ce_o(Ce^n), \\
 c_m^{r,n} &= \zeta(W_{rh}h^n + b_r),
 \end{aligned} \tag{18}$$

173 where I , f and o denote the input, forget and output gate vectors respec-
 174 tively, Ce is the cell activation vector, b is the bias vector, ϱ is the activation
 175 function, W denotes the weight matrices (*e.g.* W_{ic} is the weight matrix from
 176 the input gate to the input), \odot is the element wise product of the vectors, Ce_o
 177 and Ce_i are the cell output and cell input activation functions respectively
 178 and ζ is the network output activation function.

179 After obtaining the function F_m , it can then be used to predict the POD
 180 coefficients at current time level n . The offline calculation of snapshots at
 181 the training stage and the online procedure for constructing and resolving
 182 the P-NIROM can be algorithmically summarised in Figure 1. The details
 183 of the offline and on-line calculations are further given in Algorithm 1 and 2

184 respectively.

185

Algorithm 1: Offline Calculations

- (1) Select a set of training inputs (here the emission intensity) $\mathbf{Q}_{tr,i} \in R^P$, where, $i \in (1, \dots, N_{tr})$;
 - (2) Given the input $\mathbf{Q}_{tr,i}$, generate the solution snapshots by running the high fidelity full model during the training period $[0, T_{tr}]$
 - for** $i = 1$ to N_{tr} **do**
 - for** $n = 1$ to T_{tr} **do**
 - Solving Equation (2):
 - $\mathcal{M}(\mathbf{Q}_{tr,i})\mathbf{c}^n = \mathbf{s}^n(\mathbf{Q}_{tr,i}, \mathbf{c}^{n-1})$
 - endfor**
 - endfor**
 - (3) Calculate the POD basis functions using SVD.
-

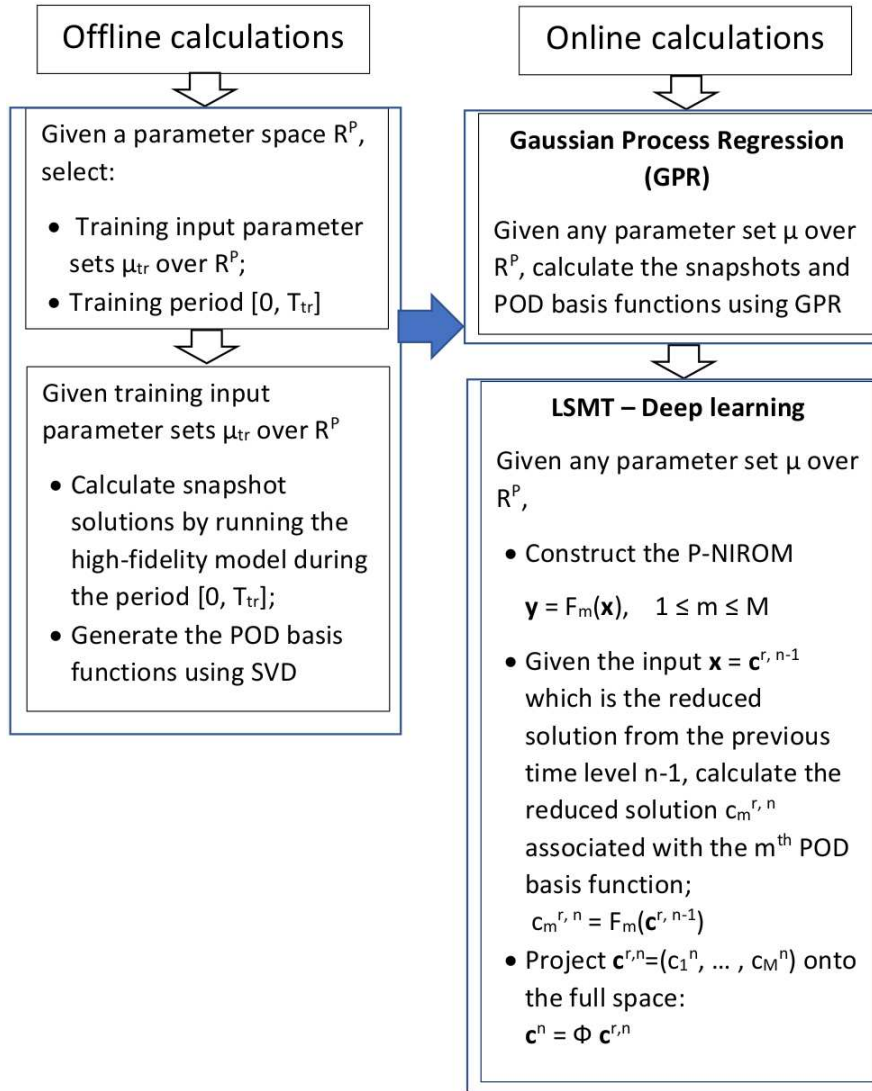


Figure 1: The figure displayed above shows the online and offline procedures of constructing and resolving the P-NIROM for any given parameter $\mu \in \mathcal{R}^P$.

Algorithm 2: Online Calculation

- (1) Using GPR, for any given unseen input $\mu = \mathbf{Q}$ (here the pollutant emission intensity), calculate the snapshot solutions
 - (i) Calculate the covariance matrices between the given unseen (\mathbf{Q}) and training (\mathbf{Q}_{tr}) points in (12) - (14);
 - (ii) For the given unseen emission intensity \mathbf{Q} , calculate the probability of the prediction of the tracer variable \mathbf{c}^r in (16) over the reduced space;
 - (iii) Calculate the best estimate of solutions over the reduced space using (17), then project back the full space.

- (2) Using LSTM, construct the set of P-NIROMs and calculate reduced solutions for any input over the parameter space;
 - (i) Using (18), construct the set of P-NIROMs F_m ($m \in (1, \dots, M)$) for the associated POD basis function Φ_m ;
 - (ii) Calculate the solutions at time level n using the P-NIROM.
 - (a) Give the reduced solution $\mathbf{c}^{r,n-1}$ at the previous time level $n - 1$;
 - (b) Calculate the solution $\mathbf{c}^{r,n} = (c_1^{r,n}, \dots, c_m^{r,n}, \dots, c_M^{r,n})$ at time level n over the reduced space

for $m = 1$ **to** M **do**

$$c_m^{r,n} = F_m(\mu, \mathbf{c}^{r,n-1})$$

endfor
 - (c) Obtain the solution $\mathbf{c}^n(\mu)$ at the current time level n by projecting $\mathbf{c}^{r,n}(\mu)$ onto the full space via:

$$\mathbf{c}^n(\mu) = \sum_{m=1}^M c_m^{r,n} \Phi_m$$

186

187 4. Regional pollutant dispersion in China

188 To demonstrate the capability of the new P-NIROM based on machine
189 learning techniques, it has been applied to a realistic case in China where
190 the SO_2 emissions from power plants disperse through the atmosphere in
191 time. The SO_2 emission intensity at the power plant locations was obtained
192 from the Regional Emission inventory in ASia (REAS 2.1) data developed
193 by National Institute of Environmental Sciences of Japan. The simulated
194 domain covers the whole Shanxi-Hebei-Shandong-Henan region of China with
195 an area encompassing $1090km \times 1060km$, and there are about 100 power
196 plants in this area [37].

197 Using adaptive mesh techniques, the 2D top adaptive mesh (20 km above
198 the sea level) is first constructed to ensure a high resolution of 2.5 km around
199 the power plant points within a radius of 6 km. The 3D unstructured mesh
200 with 61479 nodes is then obtained by extending the 2D top mesh onto the
201 terrain surface, with 11 terrain-following layers, where 7 vertical layers are
202 within 1 km above the terrain. The pollutant SO_2 sources around the power
203 plants are released into the atmosphere at the height of 200 m above the
204 terrain.

205 In the study, the simulation started at 00:00 UTC on the 10 January
206 2013 and ran through to the 15 January 2013. A time interval of $\Delta t = 0.5 hr$
207 was used. Assuming that the mixing layer height is 600 m and the turbulent
208 horizontal diffusivity is $100 m^2/s$ while the vertical eddy diffusivity is param-
209 eterised based on a scheme by Byun and Dennis [38]. The meteorological
210 fields are provided by the mesoscale meteorological model WRF (v3.5) [39].

211 In this case, the varying input parameter, $\mu = \mathbf{Q}$, is the emission intensity
212 of pollutant sources at locations ξ_1 , ξ_2 and ξ_3 (see Table 1). The emission
213 intensity of pollutant sources is ranged from 0 to $5000 g s^{-1}$. A set of training
214 pollutant sources $\mu_{tr} = \mathbf{Q}_{tr}$ at three locations is listed in Table 1. The
215 solution snapshots \mathbf{c}_{tr} with the training parameters were obtained by running
216 the high fidelity model (Fluidity [40]) and stored at equally spaced time
217 intervals (3 hrs) during the simulation period (5 days).

218 To illustrate the capability of the P-NIROM based on machine learning
219 techniques, an unseen test case, the emission intensity of pollutant sources
220 $\mu = \mathbf{Q} = (2400, 2400, 5000) g s^{-1}$, was given at locations ξ_1 , ξ_2 and ξ_3
221 respectively (T_1 in Table 1). Following the online procedure shown in Figure
222 1, using the GPR, the solution snapshots (the distribution of pollutants at
223 every 3 hrs) for the given unseen pollutant sources were calculated from the

Table 1: The emission intensity ($g s^{-1}$) of SO_2 at locations ξ_1 (x=540,y=752)km, ξ_2 (x=603, y=670)km and ξ_3 (x=753, y=679)km. $A_1 - A_{28}$ are the training cases while $T_1 - T_2$ are the unseen cases used for evaluating the predictive capability of the new P-NIROM.

Cases	ξ_1	ξ_2	ξ_3	cases	ξ_1	ξ_2	ξ_3	cases	ξ_1	ξ_2	ξ_3
A1	1047	1678	1160	A11	4267	2500	2500	A21	1440	1140	3570
A2	0	0	0	A12	2500	732	2500	A22	1250	1250	1250
A3	5000	5000	5000	A13	2500	4267	2500	A23	3750	3750	3750
A4	2500	2500	2500	A14	2500	2500	732	A24	0	5000	5000
A5	0	2500	2500	A15	2500	2500	4267	A25	1250	5000	5000
A6	5000	2500	2500	A16	534	589	910	A26	5000	5000	0
A7	2500	5000	2500	A17	600	639	1580	A27	5000	0	0
A8	2500	2500	0	A18	181	1061	1356	A28	2500	0	2500
A9	2500	2500	5000	A19	428	1881	329				
A10	732	2500	2500	A20	1300	1380	3000				
T1	2400	2400	5000	T2	5500	6000	6000				

224 training solutions for the selected training parameters (28 training parameter
225 sets in Table 1).

226 Figure 2 shows the singular values and a logarithmic scale of singular
227 values. From the calculation in (4), the sharp decrease of singular values
228 suggests that the first 36 leading POD basis functions can capture 99% of
229 dynamical energy within the solution snapshots. In this study, two cases of
230 6 and 36 POD basis functions were chosen to construct the P-NIROM. The
231 larger the number of POD basis functions chosen, the higher the accuracy of
232 P-NIROM. Figure 4 provides some of the first 36 leading basis functions. It
233 can be seen that the first leading basis function captures a large part of the
234 spatial distribution of pollutant concentration solutions, while the remaining
235 basis functions represent the details of pollutant distributions of different
236 regions.

237 A comparison of coefficients for the POD basis functions between using
238 the standard ROM and machine learning ROM (based on LSTM and GPR)
239 is provided in Figure 3. It is clearly seen that the POD coefficients are in
240 very close agreement with each other. Compared to the standard ROM,
241 the machine learning ROM has a wider range of application areas, espe-
242 cially where observational data is concerned, for example, data assimilation,
243 data reduction by condensing the information into the required dynamical

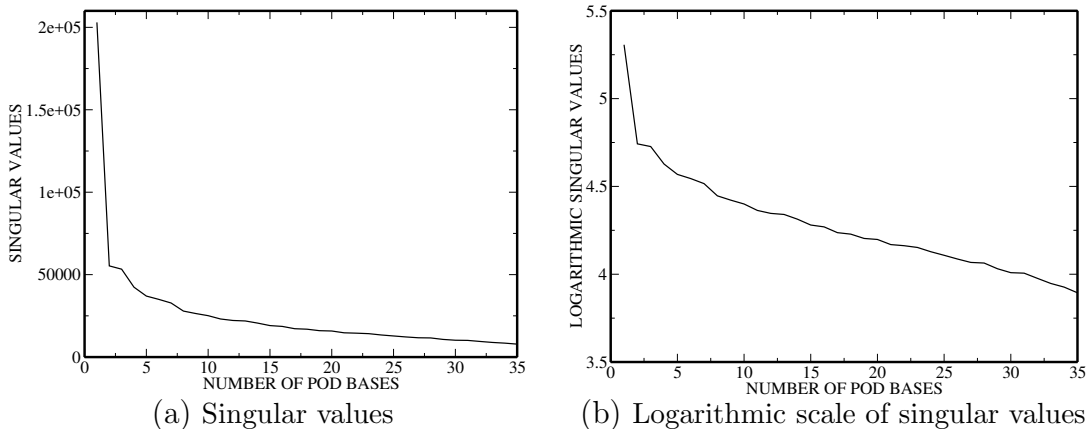


Figure 2: The singular values and logarithmic scale of singular values.

244 features.

245 Figure 5 presents the spatial distribution of pollutant solutions at time
 246 levels $t = 30 \text{ hrs}$ and $t = 105 \text{ hrs}$, as calculated by the fidelity model and P-
 247 NIROM with 6 and 36 POD basis functions. It is illustrated that P-NIROM
 248 with 6 POD basis captures most of the details of pollutant distribution at
 249 time level $t = 105 \text{ hrs}$, but fails at time level $t = 30 \text{ hrs}$. With an increased
 250 number of 36 POD basis functions, the P-NIROM has performed well at re-
 251 solving the flow dynamics and evolution of power plant plumes (see Figure
 252 5(e) and (d)). This is further highlighted in Figure 6 which shows the solu-
 253 tions from different angles. Further comparison is provided in Figure 7 which
 254 illustrates the evolution of pollutant concentrations predicted by the fidelity
 255 model and P-NIROM at the location $(x = 379, y = 786) \text{ km}$. We can see
 256 that the P-NIROM with 6 and 36 POD basis function is in close agreement
 257 with the high fidelity model at this location.

258 An error analysis of P-NIROM has been carried out. Visual inspection of
 259 Figure 8 shows the spatial distribution of absolute errors of pollutant so-
 260 lutions between the high fidelity model and P-NIROM. It is visually evident
 261 that the accuracy of P-NIROM solutions is improved by increasing the num-
 262 ber of retained POD basis functions from 6 to 36. Figure 9 illustrates the
 263 RMSE and correlation coefficients of pollutant solutions between the high
 264 fidelity model and P-NIROM with 36 POD basis functions. The correlation
 265 coefficients achieve results above 80% – 90%. This again demonstrates that
 266 the P-NIROM is in good agreement with the high fidelity full model.

267 To further investigate the predictive ability of the P-NIROM, another

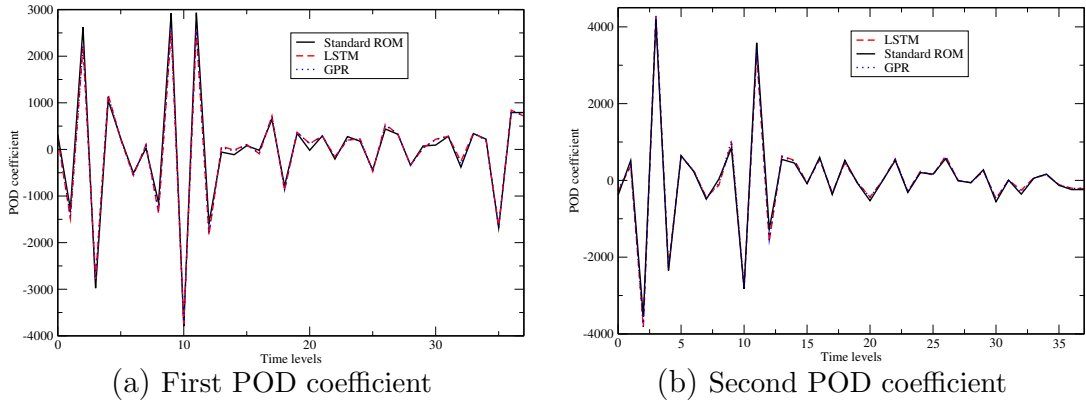
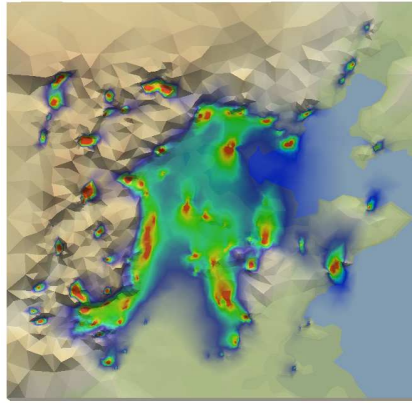
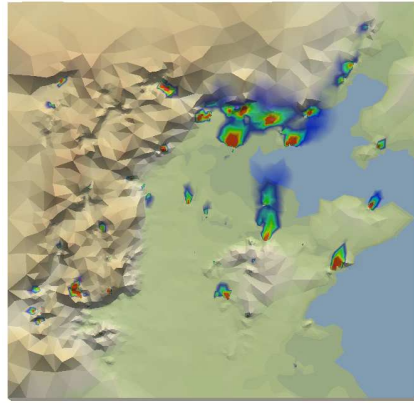


Figure 3: The first and second POD coefficients obtained from the standard ROM and machine learning ROM (the black solid line: standard ROM, the red dash line: LSTM-ROM, and the blue dot line: GPR-ROM).

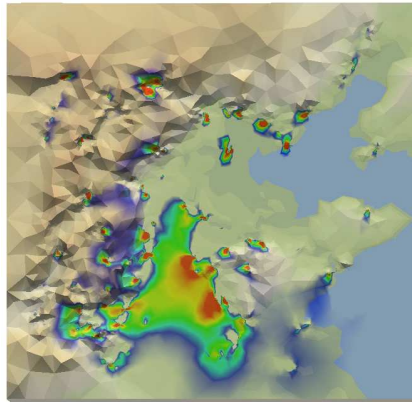
268 unseen cases (T2) was set up, where the emission intensities of pollutants
 269 at three source locations ($\mu = \mathbf{Q} = (5500, 6000, 6000) g s^{-1}$, see Table1) were
 270 given beyond the range of the training data $(0, 5000) g s^{-1}$. The pollutant
 271 solutions (at time level $t = 24 hrs$) from the high fidelity full model and
 272 P-NIROM are shown in Figures 10 (a) and (b) respectively while the cor-
 273 responding absolute error is illustrated in Figure 10 (d). A comparison of
 274 results between the high fidelity full model and P-NIROMs at a particular
 275 location $(x = 599, y = 569) km$ is provided in Figures 10 (c). As shown in
 276 the figures, the predictive ability of the P-NIROM in cases T2 is acceptable
 277 although the given test data goes beyond the range of the training data.



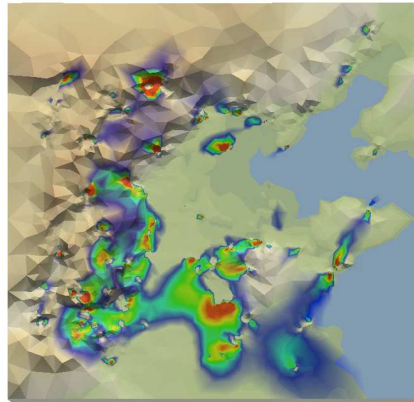
(a) 1st basis function



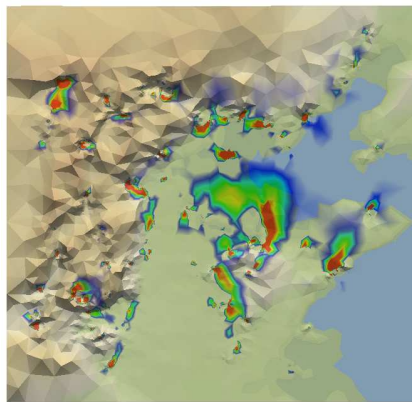
(b) 2nd basis function



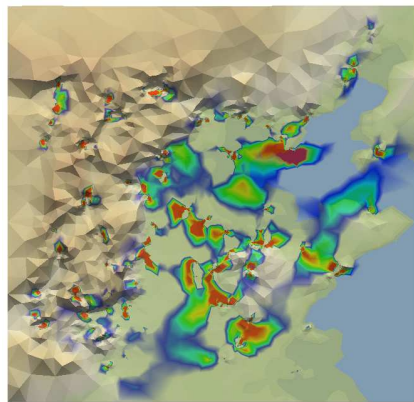
(c) 3rd basis functions



(d) 4th basis functions



(e) 31st basis functions
POD basis functions



(f) 36th basis functions
POD basis functions



Figure 4: Some of the first 36 leading POD basis functions

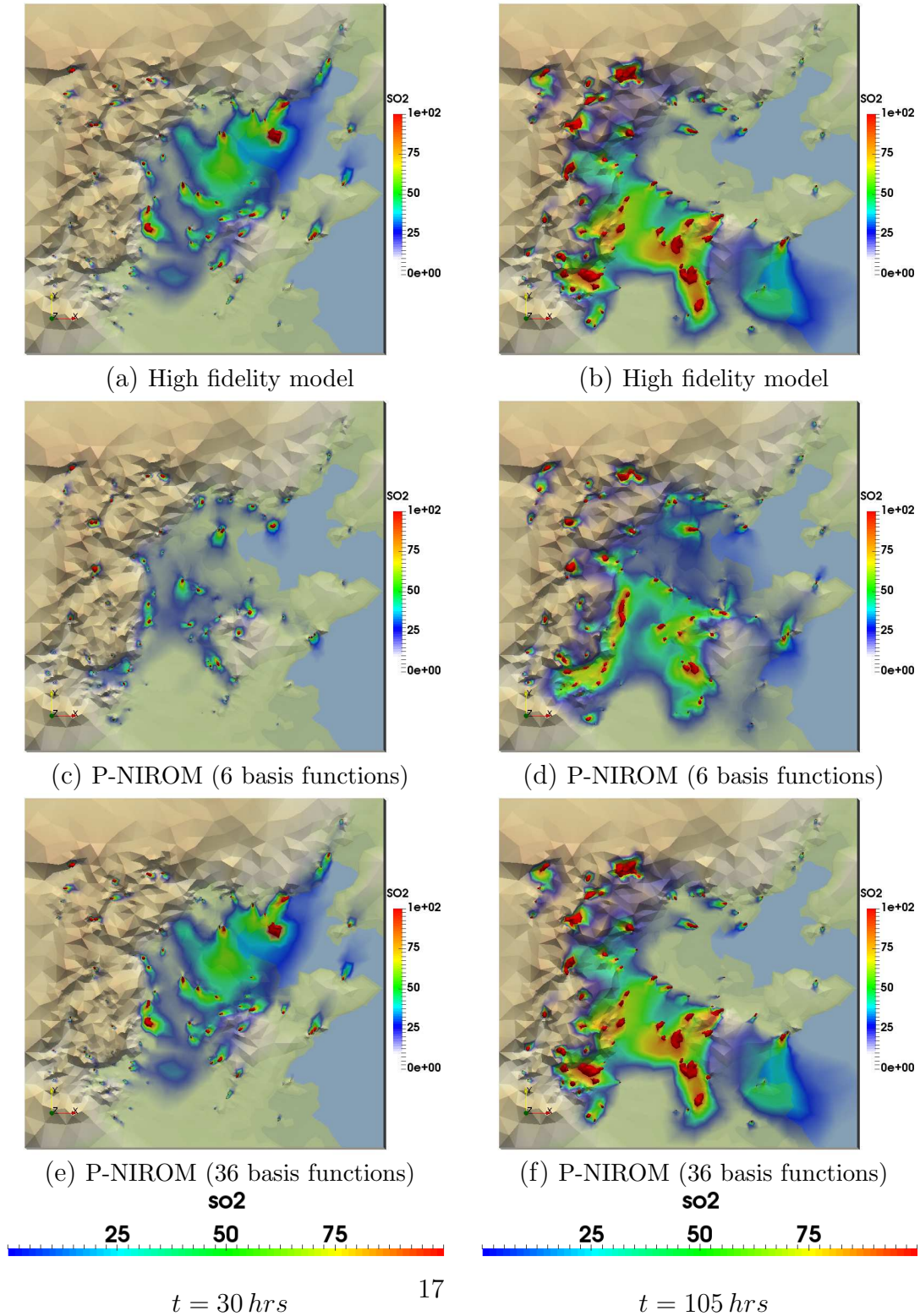
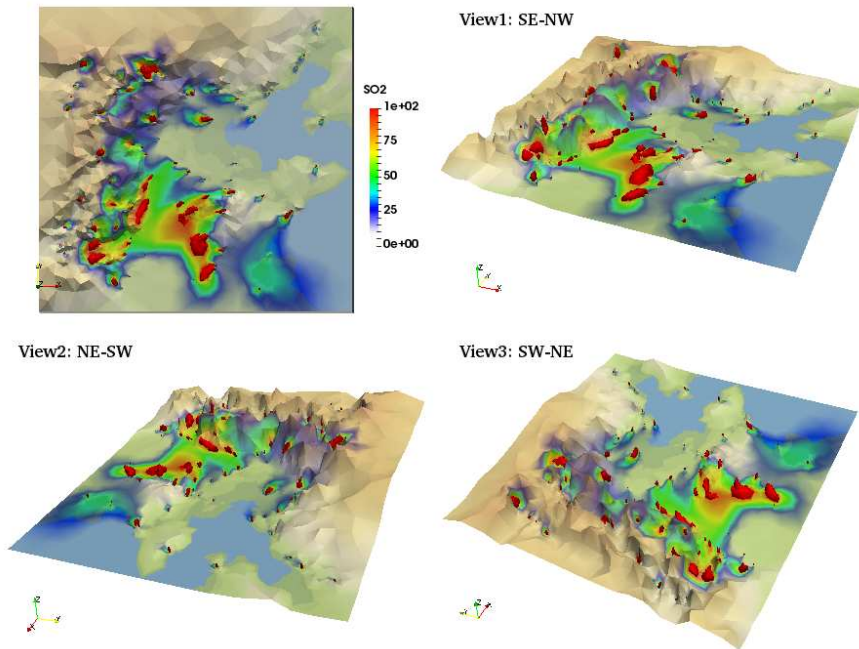
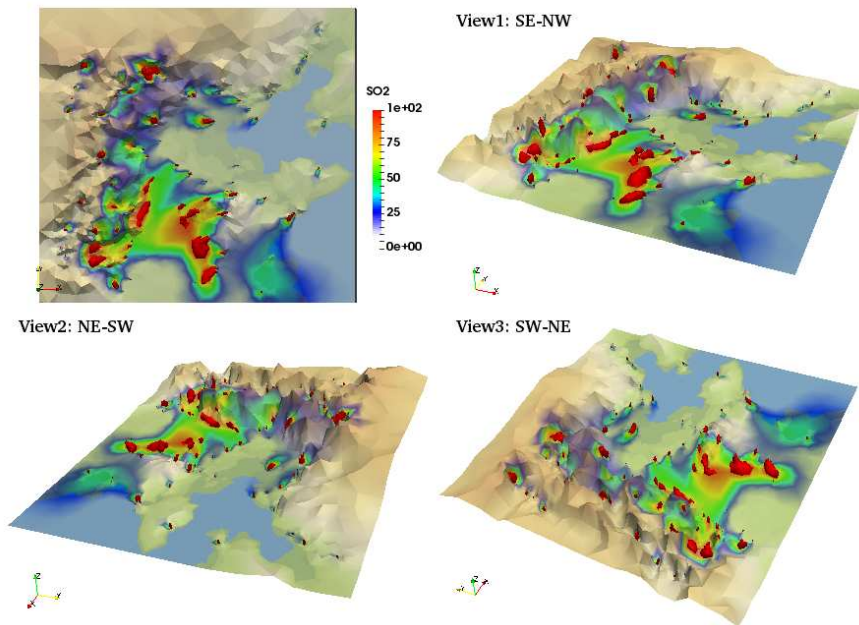


Figure 5: Case T1 ($\xi_1 = 2400, \xi_2 = 2400, \xi_3 = 5000$) $g s^{-1}$: the comparison of pollutant concentration solutions at time levels $t = 30 \text{ hrs}$ (left panel) and $t = 105 \text{ hrs}$ (right panel) between the high fidelity model and P-NIROM with 6 and 36 POD basis functions.



(a) Full model



(b) P-NIROM with 36 basis functions

Figure 6: Case T1 ($\xi_1 = 2400, \xi_2 = 2400, \xi_3 = 5000$) $g s^{-1}$: the comparison of pollutant results at time levels $t = 30$ hrs between the high fidelity full model and P-NIROM with 36 basis functions from different angles. 18

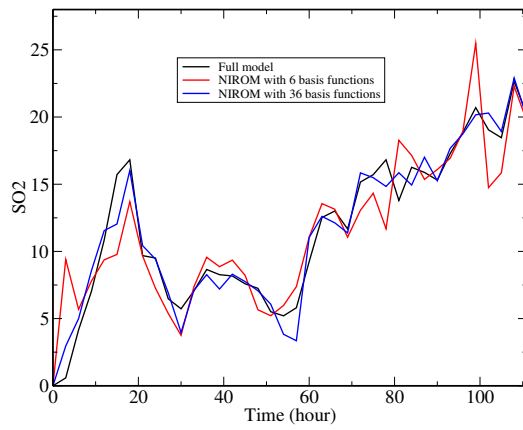
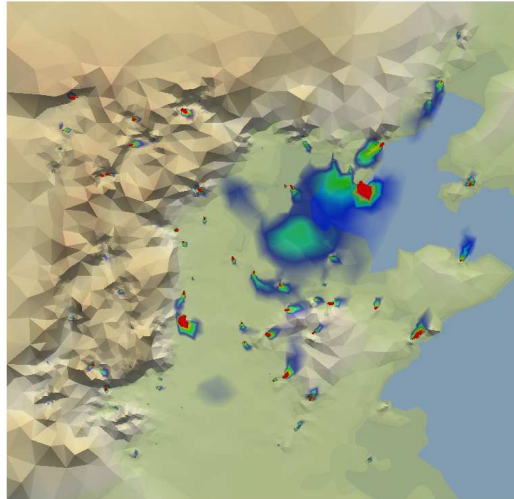
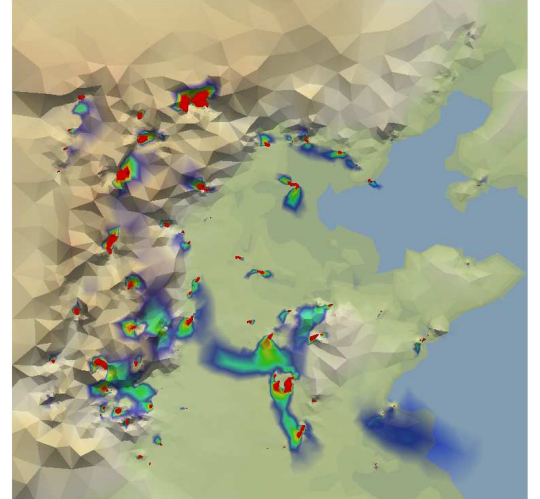


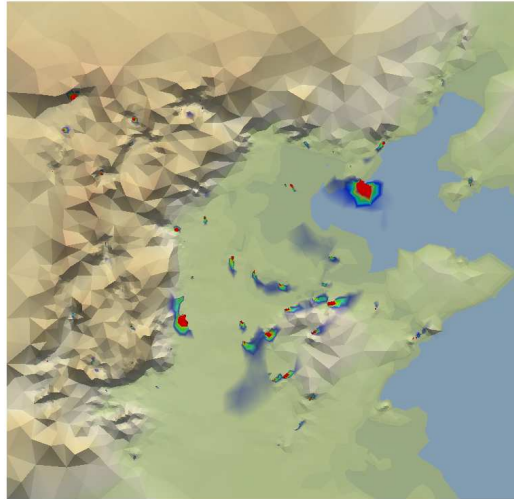
Figure 7: Case T1 ($\xi_1 = 2400, \xi_2 = 2400, \xi_3 = 5000$) $g s^{-1}$: the evolution of pollutant concentration solutions predicted by the high fidelity model and P-NIROM at a specified location $x = 378 km, y = 786 km$.



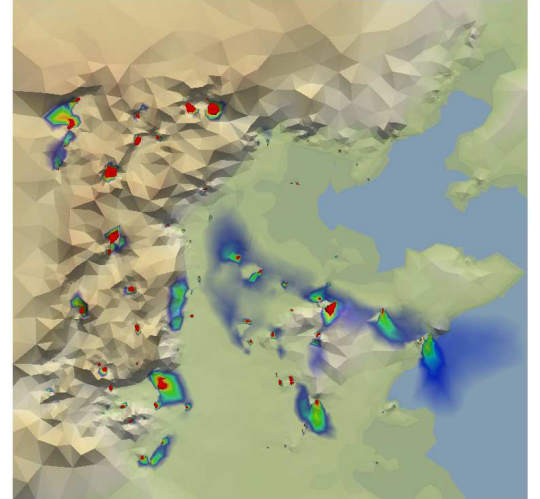
(a) Error of P-NIROM with 6 POD, $t = 30 \text{ hrs}$



(b) Error of P-NIROM with 6 POD, $t = 105 \text{ hrs}$



(c) Error of P-NIROM with 36 POD, $t = 30 \text{ hrs}$



(d) Error of P-NIROM with 36 POD, $t = 105 \text{ hrs}$

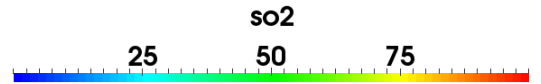
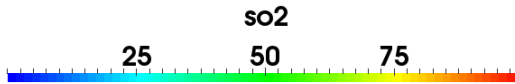


Figure 8: Case T1 ($\xi_1 = 2400, \xi_2 = 2400, \xi_3 = 5000$) $g s^{-1}$: the spatial distribution of absolute errors between the high fidelity model and P-NIROM which is constructed with 6 and 36 POD basis functions.

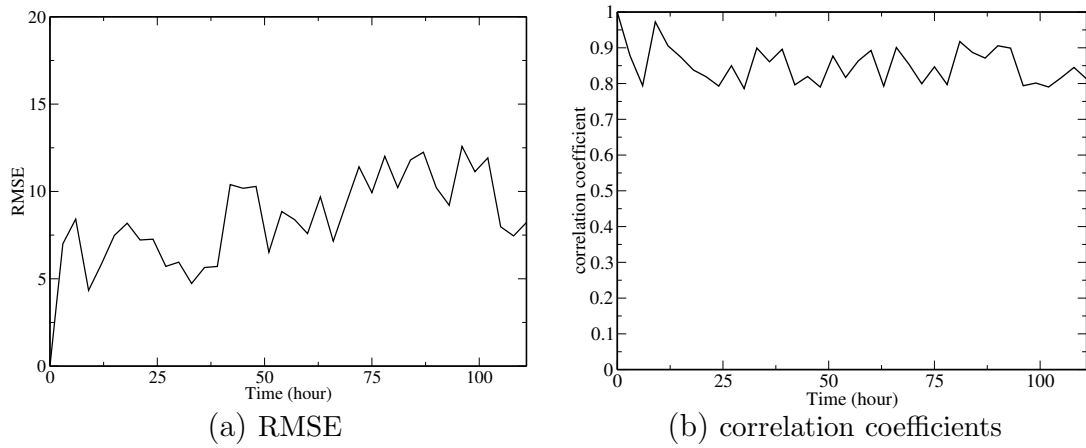
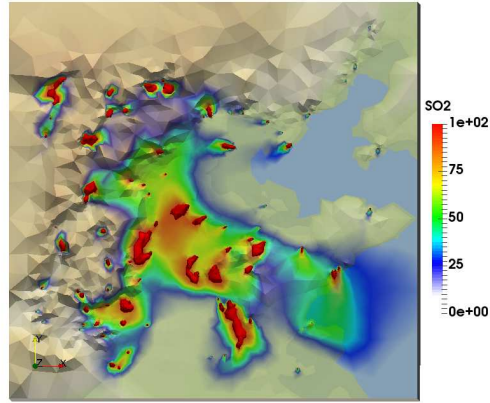
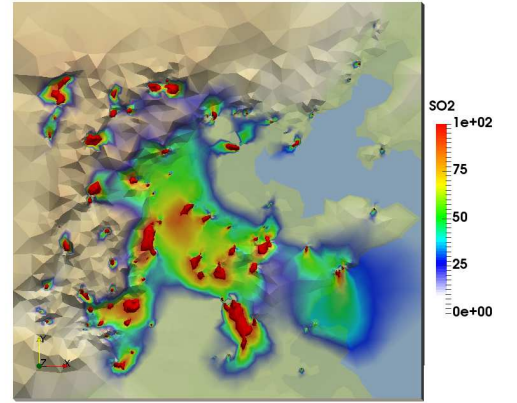


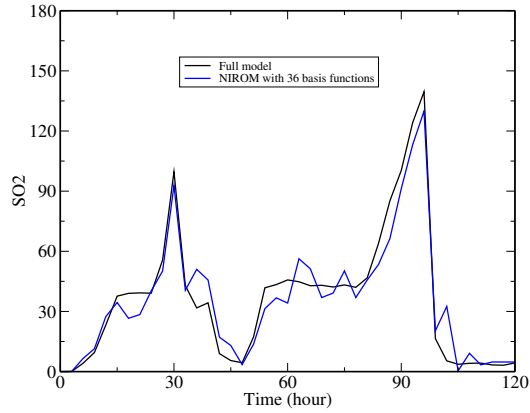
Figure 9: Case T1 ($\xi_1 = 2400, \xi_2 = 2400, \xi_3 = 5000$) $g s^{-1}$: the RMSE and correlation coefficients of pollutant concentration solutions between the high fidelity model and P-NIROM with 36 POD basis functions.



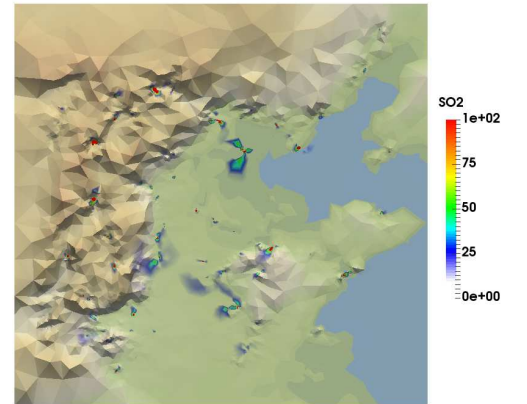
(a) High fidelity model



(b) P-NIROM



(c) Solutions at a location: $x = 599 \text{ km}$, $y = 569 \text{ km}$



(d) Error

Figure 10: Case T2 ($\xi_1 = 5500, \xi_2 = 6000, \xi_3 = 6000$) $g s^{-1}$: comparison of pollutant concentration solutions between the high fidelity full model and P-NIROM with 36 basis functions: (a) and (b) the spatial solution at time level $t = 102 \text{ hrs}$ from the high fidelity model and P-NIROM respectively; (c) the evolution of pollutant concentration solutions at a location: $x = 599 \text{ km}$, $y = 569 \text{ km}$; and (d) the spatial error at time level $t = 102 \text{ hrs}$.

278 *4.1. Computational efficiency*

279 This section provides a comparison of the online computational CPU cost
 280 required by the high fidelity full model and P-NIROM. The specifications of
 281 the machine for simulations were: 12 cores with a frequency of 3.33GHz
 282 (*Intel*[®] Xeon(R) CPU X5680 @3.33GHz × 12) and a 62.9GB memory. One
 283 core was used for the simulations since the cases were simulated in serial.
 284 Table 2 lists the online CPU cost required for running the high fidelity model
 285 and P-NIROM. The offline cost (see Figure 1) at the training stage is not
 286 listed in this table. It can be seen that using the P-NIROM, the CPU time is
 287 reduced by five order of magnitude in comparison to the high fidelity model.

Table 2: Online CPU cost required for running the high fidelity model and P-NIROM during one time step.

Cases	Model	assembling and solving	projection	interpolation	total
Test case	Full model	616.9	0	0	696.59
	NIROM	0	0.003	0.001	0.004

288

289 **5. Conclusions**

290 This article has presented a new P-NIROM for predictive modelling of
 291 pollutant transport phenomena. The machine learning techniques in com-
 292 bination with POD are used for constructing the P-NIROM. First, at the
 293 training stage, for the selected input parameters $\mu_{tr} \in R^P$, the solution snap-
 294 shots (serving as training datasets) and POD basis functions are obtained
 295 by running the high fidelity model. From the training data sets, for any
 296 given input parameters $\mu \in R^P$, using the machine learning technique a set
 297 of hyper-surface functions (P-NIROMs) is constructed to represent the dy-
 298 namics of pollutant transport over the reduced space. The P-NIROM is then
 299 used for calculating the reduced solutions (POD coefficients) for the given μ
 300 (the emission intensity). The unique combination of the P-NIROM and ma-
 301 chine learning techniques enables rapid and reasonably accurate simulations.
 302 The P-NIROM techniques developed here are robust and can be used for a
 303 large number of disciplines not least of pollutant flow based disciplines.

304 The P-NIROM has been applied to a realistic case in China involving
 305 plumes released from over 100 power plants. The varying input parameter is

306 the emission intensity of pollutant sources. A comparison of pollutant solu-
307 tions between the high fidelity model and P-NIROM has been undertaken.
308 The P-NIROM with 36 POD basis functions exhibits an overall good agree-
309 ment with the high fidelity model. The online computation cost required by
310 the P-NIROM is reduced by several orders of magnitude in comparison to
311 the high fidelity model.

312 Compared to existing P-NIROM techniques (for example, based on ra-
313 dial basis functions), the P-NIROM based on machine learning methods pro-
314 vides a wider range of application areas, for example, uncertainty analysis in
315 both data and modelling results, real-time interactive use, data management
316 (real-time data monitoring/analysis), data assimilation and better-informed
317 decision making. In particular, the machine learning techniques with ROM
318 can be used for data selection and data reduction by condensing the infor-
319 mation into the desired number of features and recovering the original data
320 from the reduced feature set.

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324 References

- 325 [1] Kristen M Foley, Sergey L Napelenok, Carey Jang, Sharon Phillips,
326 Bryan J Hubbell, and Charles M Fulcher. Two reduced form air quality
327 modeling techniques for rapidly calculating pollutant mitigation poten-
328 tial across many sources, locations and precursor emission types. *Atmo-
329 spheric environment*, 98:283–289, 2014.
- 330 [2] Laurent Cordier, Bernd R Noack, Gilles Tissot, Guillaume Lehnasch,
331 Joël Delville, Maciej Balajewicz, Guillaume Daviller, and Robert K
332 Niven. Identification strategies for model-based control. *Experiments
333 in fluids*, 54(8):1580, 2013.
- 334 [3] Bernard Haasdonk. Reduced basis methods for parametrized pdesa tu-
335 torial introduction for stationary and instationary problems. *Model re-
336 duction and approximation: theory and algorithms*, 15:65, 2017.

- 337 [4] Peter Benner, Serkan Gugercin, and Karen Willcox. A survey of
338 projection-based model reduction methods for parametric dynamical
339 systems. *SIAM Review*, 57(4):483–531, 2015.
- 340 [5] Michael Hinze and Stefan Volkwein. Proper orthogonal decomposition
341 surrogate models for nonlinear dynamical systems: Error estimates and
342 suboptimal control. In *Dimension reduction of large-scale systems*, pages
343 261–306. Springer, 2005.
- 344 [6] Rafik Djouad and Bruno Sportisse. Solving reduced chemical models in
345 air pollution modelling. *Applied Numerical Mathematics*, 44(1-2):49–61,
346 2003.
- 347 [7] Janelle K HAMMOND, Rachida CHAKIR, Frédéric Bourquin, and
348 Yvon Maday. PBDW: a non-intrusive Reduced Basis Data Assimilation
349 Method and its application to outdoor Air Quality Models. working
350 paper or preprint, January 2018.
- 351 [8] Farha A Alkuwari, Serge Guillas, and Yuhang Wang. Statistical down-
352 scaling of an air quality model using fitted empirical orthogonal func-
353 tions. *Atmospheric environment*, 81:1–10, 2013.
- 354 [9] F. Fang, T. Zhang, D. Pavlidis, C.C. Pain, A.G. Buchan, and I.M.
355 Navon. Reduced order modelling of an unstructured mesh air pollution
356 model and application in 2D/3D urban street canyons. *Atmospheric
357 Environment*, 96:96–106, 2014.
- 358 [10] H. Chen. Blackbox stencil interpolation method for model reduction.
359 Master’s thesis, Massachusetts Institute of Technology, 2012.
- 360 [11] Michael Schlegel and Bernd R. Noack. On long-term boundedness of
361 Galerkin models. *Journal of Fluid Mechanics*, 765:325–352, 2 2015.
- 362 [12] Jan Osth, Bernd R. Noack, Sinia Krajnovi, Diogo Barros, and Jacques
363 Bore. On the need for a nonlinear subscale turbulence term in POD
364 models as exemplified for a high-Reynolds-number flow over an Ahmed
365 body. *Journal of Fluid Mechanics*, 747:518–544, 5 2014.
- 366 [13] David Amsallem and Charbel Farhat. Stabilization of projection-based
367 reduced-order models. *International Journal for Numerical Methods in
368 Engineering*, 91(4):358–377, 2012.

- 369 [14] Leopoldo P. Franca and Sergio L. Frey. Stabilized finite element meth-
370 ods: II. The incompressible Navier-Stokes equations. *Computer Methods*
371 *in Applied Mechanics and Engineering*, 99(2-3):209–233, 1992.
- 372 [15] S. Chaturantabut and D.C. Sorensen. Nonlinear model reduction via
373 discrete empirical interpolation. *SIAM J. Sci. Comput*, 32:2737–2764,
374 2010.
- 375 [16] Feriedoun S. and Alireza J. α Regularization of the POD-Galerkin dy-
376 namical systems of the Kuramoto-Sivashinsky equation. *Applied Math-*
377 *ematics and Computation*, 218(10):6012 – 6026, 2012.
- 378 [17] D. Xiao, F. Fang, A.G. Buchan, C.C. Pain, I.M. Navon, J. Du, , and
379 G. Hu. Non-linear model reduction for the Navier-Stokes equations using
380 Residual DEIM method. *Journal of Computational Physics*, 263:1–18,
381 2014.
- 382 [18] D. Xiao, F. Fang, J. Du, C.C. Pain, I.M. Navon, A.G. Buchan, A.H.
383 ElSheikh, and G. Hu. Non-linear Petrov-Galerkin methods for reduced
384 order modelling of the Navier-Stokes equations using a mixed finite el-
385 ement pair. *Computer Methods In Applied Mechanics and Engineering*,
386 255:147–157, 2013.
- 387 [19] C. Audouze, F.D. Vuyst, and P.B. Nair. Nonintrusive reduced-order
388 modeling of parametrized time-dependent partial differential equations.
389 *Numerical Methods for Partial Differential Equations*, 29(5):1587–1628,
390 2013.
- 391 [20] J.S. Hesthaven and S. Ubbiali. Non-intrusive reduced order modeling
392 of nonlinear problems using neural networks. *Journal of Computational*
393 *Physics*, 363:55 – 78, 2018.
- 394 [21] D. Xiao, F. Fang, A.G. Buchan, C.C. Pain, I.M. Navon, and A. Mug-
395 geridge. Non-intrusive reduced order modelling of the Navier–Stokes
396 equations. *Computer Methods in Applied Mechanics and Engineering*,
397 293:552–541, 2015.
- 398 [22] Wang Chen, Jan S Hesthaven, Bai Junqiang, Zhang Yang, and Yang
399 Tihao. A greedy non-intrusive reduced order model for fluid dynamics.
400 *Journal of Northwestern Polytechnical University*, 2017.

- 401 [23] Diana Alina Bistrián and Ionel Michael Navon. Randomized dynamic
402 mode decomposition for nonintrusive reduced order modelling. *Inter-*
403 *national Journal for Numerical Methods in Engineering*, 112(1):3–25,
404 2017.
- 405 [24] D. Xiao, F. Fang, C. Pain, and G. Hu. Non-intrusive reduced order
406 modelling of the Navier-Stokes equations based on RBF interpolation.
407 *International Journal for Numerical Methods in Fluids*, 79(11):580–595,
408 2015.
- 409 [25] D Xiao, F Fang, C Pain, IM Navon, and A Muggeridge. Non-intrusive
410 Reduced Order Modelling of Waterflooding in Geologically Heteroge-
411 neous Reservoirs. In *ECMOR XV-15th European Conference on the*
412 *Mathematics of Oil Recovery*, 2016.
- 413 [26] Toomaj Foroud and Abbas Seifi. A guided pattern search with a
414 non-intrusive reduced order modeling for oil production optimization:
415 Brugge field case study. *Journal of Petroleum Science and Engineering*,
416 147:570–584, 2016.
- 417 [27] D Xiao, Z Lin, F Fang, C C Pain, Ionel M Navon, P Salinas, and
418 A Muggeridge. Non-intrusive reduced-order modeling for multiphase
419 porous media flows using smolyak sparse grids. *International Journal*
420 *for Numerical Methods in Fluids*, 83(2):205–219, 2017.
- 421 [28] D. Xiao, P. Yang, F. Fang, J. Xiang, C.C. Pain, and I.M. Navon. Non-
422 intrusive reduced order modeling of fluid-structure interactions. *Com-*
423 *puter Methods in Applied Mechanics and Engineering*, 303:35–54, 2016.
- 424 [29] Z. Wang, D. Xiao, F. Fang, R. Govindan, C.C. Pain, and Y. Guo. Model
425 identification of reduced order fluid dynamics systems using deep learn-
426 ing. *International Journal for Numerical Methods in Fluids*, 86(4):255–
427 268, 2018.
- 428 [30] Yann LeCun, Yoshua Bengio, and Geoffrey Hinton. Deep learning. *Na-*
429 *ture*, 521(7553):436–444, 2015.
- 430 [31] Geoffrey Hinton, Li Deng, Dong Yu, George E Dahl, Abdel-rahman
431 Mohamed, Navdeep Jaitly, Andrew Senior, Vincent Vanhoucke, Patrick

- 432 Nguyen, Tara N Sainath, et al. Deep neural networks for acoustic mod-
433 eling in speech recognition: The shared views of four research groups.
434 *IEEE Signal Processing Magazine*, 29(6):82–97, 2012.
- 435 [32] Jonathan J Tompson, Arjun Jain, Yann LeCun, and Christoph Bregler.
436 Joint training of a convolutional network and a graphical model for hu-
437 man pose estimation. In *28th Annual Conference on Neural Information*
438 *Processing Systems 2014, NIPS 2014 - Montreal, Canada, Advances in*
439 *neural information processing systems*, pages 1799–1807, 2014.
- 440 [33] Michael KK Leung, Hui Yuan Xiong, Leo J Lee, and Brendan J Frey.
441 Deep learning of the tissue-regulated splicing code. *Bioinformatics*,
442 30(12):i121–i129, 2014.
- 443 [34] Raia Hadsell, Pierre Sermanet, Jan Ben, Ayse Erkan, Marco Scoffier,
444 Koray Kavukcuoglu, Urs Muller, and Yann LeCun. Learning long-
445 range vision for autonomous off-road driving. *Journal of Field Robotics*,
446 26(2):120–144, 2009.
- 447 [35] Ronan Collobert, Jason Weston, Léon Bottou, Michael Karlen, Koray
448 Kavukcuoglu, and Pavel Kuksa. Natural language processing (almost)
449 from scratch. *Journal of Machine Learning Research*, 12(Aug):2493–
450 2537, 2011.
- 451 [36] Carl Edward Rasmussen. *Gaussian Processes in Machine Learning*,
452 pages 63–71. Springer Berlin Heidelberg, Berlin, Heidelberg, 2004.
- 453 [37] J Zheng, J Zhu, Z Wang, F Fang, CC Pain, and J Xiang. Towards
454 a new multiscale air quality transport model using the fully unstruc-
455 tured anisotropic adaptive mesh technology of fluidity (version 4.1. 9).
456 *Geoscientific Model Development*, 8(10):3421, 2015.
- 457 [38] Daewon W Byun and Robin Dennis. Design artifacts in eulerian air
458 quality models: Evaluation of the effects of layer thickness and vertical
459 profile correction on surface ozone concentrations. *Atmospheric Envi-*
460 *ronment*, 29(1):105–126, 1995.
- 461 [39] William C Skamarock, Joseph B Klemp, Jimy Dudhia, David O Gill,
462 Dale M Barker, Wei Wang, and Jordan G Powers. A description of the
463 advanced research wrf version 3. Technical report, NCAR Tech, June
464 2008.

465 [40] AMCG and Imperial. College. London. Fluidity manual v4.1.12.
466 figshare. <https://dx.doi.org/10.6084/m9.figshare.1387713.v2>, pages 1–
467 329, 2015.